

Calorimetrically determined U(VI) toxicity in Brassica napus correlates with oxidoreductase activity and U(VI) speciation

Sachs, S.; Geipel, G.; Bok, F.; Oertel, J.; Fahmy, K.;

Originally published:

August 2017

Environmental Science & Technology 51(2017)18, 10843-10849

DOI: <https://doi.org/10.1021/acs.est.7b02564>

Perma-Link to Publication Repository of HZDR:

<https://www.hzdr.de/publications/Publ-25562>

Release of the secondary publication
on the basis of the German Copyright Law § 38 Section 4.

Calorimetrically determined U(VI) toxicity in *Brassica napus* correlates with oxidoreductase activity and U(VI) speciation

Susanne Sachs, Gerhard Geipel, Frank Bok, Jana Oertel, and Karim Fahmy

Environ. Sci. Technol., **Just Accepted Manuscript** • DOI: 10.1021/acs.est.7b02564 • Publication Date (Web): 25 Aug 2017

Downloaded from <http://pubs.acs.org> on September 1, 2017

Just Accepted

“Just Accepted” manuscripts have been peer-reviewed and accepted for publication. They are posted online prior to technical editing, formatting for publication and author proofing. The American Chemical Society provides “Just Accepted” as a free service to the research community to expedite the dissemination of scientific material as soon as possible after acceptance. “Just Accepted” manuscripts appear in full in PDF format accompanied by an HTML abstract. “Just Accepted” manuscripts have been fully peer reviewed, but should not be considered the official version of record. They are accessible to all readers and citable by the Digital Object Identifier (DOI®). “Just Accepted” is an optional service offered to authors. Therefore, the “Just Accepted” Web site may not include all articles that will be published in the journal. After a manuscript is technically edited and formatted, it will be removed from the “Just Accepted” Web site and published as an ASAP article. Note that technical editing may introduce minor changes to the manuscript text and/or graphics which could affect content, and all legal disclaimers and ethical guidelines that apply to the journal pertain. ACS cannot be held responsible for errors or consequences arising from the use of information contained in these “Just Accepted” manuscripts.

**Calorimetrically determined U(VI) toxicity in *Brassica napus*
correlates with oxidoreductase activity and U(VI) speciation**

Susanne Sachs*, Gerhard Geipel, Frank Bok, Jana Oertel*, Karim Fahmy*

Helmholtz-Zentrum Dresden - Rossendorf, Institute of Resource Ecology,
Bautzner Landstraße 400, 01328 Dresden, Germany

*corresponding author

k.fahmy@hzdr.de

1 ABSTRACT

2 Radioecological studies depend on the quantitative toxicity assessment of environmental
3 radionuclides. At low dose exposure, the life span of affected organisms is barely shortened
4 enabling the transfer of radionuclides through an almost intact food chain. Lethality-based
5 toxicity estimates are not adequate in this regime because they require higher
6 concentrations. However, increased radionuclide concentration alters its speciation,
7 rendering the extrapolation to the low dose exposure chemically inconsistent. Here, we
8 demonstrate that microcalorimetry provides a sensitive real-time monitor of toxicity of
9 uranium (in the U(VI) oxidation state) in a plant cell model of *Brassica napus*. We introduce
10 the calorimetric descriptor “metabolic capacity” and show that it correlates with
11 enzymatically determined cell viability. It is independent of physiological models and robust
12 against the naturally occurring fluctuations in the metabolic response to U(VI) of plant cell
13 cultures. In combination with time-resolved laser-induced fluorescence spectroscopy and
14 thermodynamic modeling, we show that the plant cell metabolism is affected predominantly
15 by hydroxo-species of U(VI) with an IC₅₀ threshold of ~90 μM. The data emphasize the yet
16 little exploited potential of microcalorimetry for the speciation-sensitive ecotoxicology of
17 radionuclides.

18

19 Keywords

20 Uranium, plant cells, metabolism, isothermal microcalorimetry, speciation, TRIFS,
21 thermodynamic modeling

22

23 INTRODUCTION

24 The transfer of environmental radionuclides into the food chain is a central concern in the
25 safety assessment of both nuclear waste repositories and remediation strategies in
26 radioactively contaminated areas. The interaction of radionuclides with plants is mostly
27 described by transfer factors without knowledge of underlying mechanisms. However,
28 recent studies of the interaction of radionuclides, e.g., uranium, with plants revealed the
29 importance of radionuclide speciation. The latter was correlated with uranium uptake from
30 nutrient medium and translocation in plants.¹ Cross-species studies showed that both
31 processes are speciation-dependent^{2,3} and uranium toxicity can be further modulated by
32 phosphate.⁴ In addition to speciation effects on uranium uptake and oxidative stress
33 response,^{5,6} the redox state of uranium and the intracellular glutathione pool have also
34 been investigated in plants.⁷ The *in situ* speciation of uranium in plants⁸ and their subcellular
35 compartments⁹ has been observed by spectroscopy. In response to heavy metal stress,
36 plants synthesize protective metal-binding metabolites, store metal chelates in the vacuole
37 or secrete them into the rhizosphere.¹⁰ This reduces metal toxicity which originates in the
38 replacement of natural metal cofactors from enzymes, the functional inhibition of sulfhydryl
39 group-containing proteins or the accelerated formation of reactive oxygen species.¹⁰

40 Correlating molecular information on radionuclide speciation and biomolecular
41 interactions with physiological performance is a major challenge for quantitative
42 radioecology. We have recently shown that microcalorimetric monitoring of metabolic
43 activity in combination with genetic engineering can identify molecular details of the
44 modulation of uranium toxicity in a living microorganism.¹¹ However, the application of
45 microcalorimetry to plant metabolism poses unique challenges. In contrast to bacteria which
46 exhibit exponential growth phases from which division rates can be derived as

47 physiologically meaningful parameters of toxicity, this is not the case with plant cells. Their
48 metabolism does not follow simple mathematical models and typically declines under the
49 conditions of calorimetric measurements, where photosynthesis is not supported. Whereas
50 non-photosynthetic experimental conditions are compatible with plant cell culture, there is a
51 lack of model-free descriptors that can be used to derive quantitative measures of toxicity
52 from metabolic monitoring. Such descriptors would greatly enhance the value of
53 microcalorimetry in radioecology, because metabolic monitoring has reached a degree of
54 sensitivity that allows detecting actinide toxicity in the environmentally relevant
55 concentrations. Here, lethality is negligible and thus inadequate to derive realistic toxicity
56 measures. In contrast, metabolic responses are clearly visible and render themselves the
57 most sensitive and also biologically most meaningful sensor of toxicity.

58 In order to overcome these restrictions, we have empirically determined a model-
59 independent descriptor, i.e., “metabolic capacity”, that allows evaluating calorimetric data
60 of declining metabolic phases as typically found with plant cells. In the present work, we
61 used this approach to investigate the concentration-dependent influence of U(VI) on plant
62 cell metabolism using canola callus cells (*Brassica napus*). *B. napus* is known to be able to
63 accumulate heavy metals in higher quantities than many other species.³ The callus cells are
64 superior over non-callus cells due to their simpler organization and the better control of
65 their growth conditions. At the same time, they retain the ability to synthesize typical
66 secondary metabolites of intact tissues,¹² which renders them a suitable model system for
67 metabolic studies. The “metabolic capacity” of the cells in the presence of 0 to 200 μM U(VI)
68 was determined and correlated with cell viability. We were particularly interested in relating
69 these data to uranium speciation. Therefore, the latter was further assessed by time-
70 resolved laser-induced fluorescence spectroscopy and thermodynamic modeling. Our data

71 provide both a physiological validation of the descriptor “metabolic capacity” and a
72 speciation-dependent quantification of U(VI) toxicity in a well-established plant cell model.

73

74 **MATERIALS AND METHODS**

75 **Cell Cultivation.** Callus cells from *B. napus* (PC-1113) were obtained from DSMZ
76 (Braunschweig, Germany). Friable cells were transferred to liquid modified Linsmaier and
77 Skoog medium (medium R¹³) to initiate growth of suspension cell cultures on an orbital
78 shaker at room temperature. After a 7 days growth cycle the cells were subcultured into
79 fresh culture medium to maintain the suspension culture. Cell cultures from passage number
80 2 – 11 were used for microcalorimetry and viability measurements.

81 **Microcalorimetry.** Cell suspensions (20 mL) were filtered through a nylon mesh (50 µm
82 pore size) without suction. Subsequently, the cells were rinsed with 10 mL medium R with a
83 reduced phosphate concentration of 12.5 µM representing 1% of the original phosphate
84 concentration (medium R_{red}; pH 5.8; Table S1, supporting information) in order to minimize
85 precipitation of U(VI) phosphate complexes in experiments with uranium. Isothermal
86 calorimetric measurements were performed with a TAM III (Thermal Activity Monitor)
87 instrument (Waters GmbH, Eschborn, Germany) equipped with 12 microcalorimeters in twin
88 configuration (one side for the sample the other for an aluminium reference). Sample
89 preparation time was kept as short as possible. Wet cells (0.3 g) were transferred into 4 mL
90 ampoules with 2 mL of medium R_{red}. A U(VI) stock solution was prepared by dissolution of
91 UO₂(NO₃)₂ × 6 H₂O in Milli-Q water (Milli-RO/Milli-Q-System, Millipore, Molsheim, France)
92 and subsequent filtration through 0.2 µm filters (Filtropur S, Sarstedt, Nümbrecht,
93 Germany). The final concentration was 9.58 mM UO₂(NO₃)₂ as determined by inductively
94 coupled plasma-mass spectrometry (ICP-MS; model ELAN 9000, Perkin Elmer, Boston, USA).

95 Aliquots of this stock solution were added to the cell suspensions giving final concentrations
96 of 20-200 μM U(VI). The ampoules were tightly capped, including control samples of
97 medium R_{red} and of cells in medium R_{red} in the absence of U(VI). The samples were held in
98 the TAM III in a waiting position for 15 min before complete insertion followed by 45 min
99 equilibration. In each experiment, thermograms were recorded at least in duplicates in the
100 absence and in the presence of different U(VI) concentrations. Seven independent
101 microcalorimetric experiments with different cell passages were evaluated (details are given
102 in Fig. 2). Over the course of the experiments (up to 300 hours), the pH stayed between 5.5
103 and 5.9.

104 **Viability Measurements.** Cell viability was measured by the MTT test.¹⁴ It detects the
105 activity of mitochondrial and cytosolic dehydrogenases which reduce water-soluble 3-(4,5-
106 dimethylthiazol-2-yl)-2,5-diphenyl-tetrazolium bromide (MTT; Duchefa, Harlem, The
107 Netherlands) to the water-insoluble formazan product accompanied by a yellow to blue
108 color change.¹⁵ A stock solution of MTT (5 mg/mL) was prepared in phosphate buffered
109 saline (PBS, without Ca^{2+} and Mg^{2+} ; Biochrom, Berlin, Germany) and passed through a 0.2 μm
110 filter (Filtropur S). At the end of calorimetric data acquisition, the pH values of the nutrient
111 media were measured within the ampoules (inoLab pH meter pH720 and SenTix Mic pH
112 electrode, WTW, Weilheim, Germany). Subsequently, the supernatant medium was removed
113 and the cells were washed twice with 1 mL PBS solution followed by the addition of 1 mL
114 PBS and 200 μL MTT stock solution to each sample. The ampoules were sealed and
115 incubated in the dark under gentle agitation for 3 hours at room temperature. The
116 supernatants were removed and 1 mL 0.04 M HCl (p.a., Merck Darmstadt, Germany) in
117 isopropanol (p.a., Roth, Karlsruhe, Germany) was added to each cell sample to dissolve the
118 formazan crystals. After slight agitation for 10 min at room temperature, 8 \times 100 μL of each

119 isopropanol solution was pipetted into 96-well plates (CELLSTAR®, Greiner Bio-one,
120 Frickenhausen, Germany) and the absorbance at 620 nm determined in a microplate reader
121 (Mithras LB940, Berthold, Bad Wildbad, Germany). Viability of U(VI)-exposed cells was
122 expressed as the absorbance in percent of that from non-exposed control samples. The
123 results represent mean values and standard errors of the mean of a total of 12-24
124 independent samples.

125 **Time-Resolved Laser-Induced Fluorescence Measurements.** Time-resolved laser-induced
126 fluorescence (TRLFS) measurements were performed under ambient conditions at room
127 temperature in order to characterize the U(VI) speciation in medium R_{red} . Aliquots of the
128 9.58 mM $UO_2(NO_3)_2$ stock solution were added to medium R_{red} to give final concentrations of
129 20-200 μ M U(VI). The pH values of these solutions were readjusted to pH 5.8 with diluted
130 NaOH (Merck) solutions, if necessary. TRLFS studies were performed using a Nd:YAG
131 pumped OPO system (New-Port Spectra Physics, Quanta Ray, USA) with a repetition rate of
132 20 Hz and laser energies of about 4.4 mJ.⁹ The excitation wavelength was 440 nm. Emission
133 signals were focused on the entrance slit of a 270 mm spectrograph (SP2300, Acton
134 Research, Roper Scientific, Martinsried, Germany) and the luminescence spectra detected
135 with an intensified camera system (PIMAX3, Princeton Instruments, Roper Scientific). Using
136 the internal delay generator, time-resolved spectra were recorded during a gate width of
137 500 ns. In 101 delay steps the gate was shifted to delay times of 0.5 or 5 μ s with a step size
138 of 5 or 50 ns, respectively. At each delay time, spectra were recorded with 100 laser pulses
139 per spectrum in the wavelength range between 450.4 and 727.0 nm at a resolution of
140 0.266 nm. The start wavelength of about 450 nm was selected to avoid scattered light from
141 the exciting laser pulse on the camera. The emission spectra were recorded with WinSpec32

142 (Roper Scientific), converted into ASCII files and evaluated with the OriginPro 2015G
 143 software (OriginLab Corporation, Northampton, USA).

144

145 RESULTS AND DISCUSSION

146 **Isothermal Microcalorimetry.** We have used heterotrophically growing *B. napus* cells as a
 147 model system to investigate interference of U(VI) with plant metabolism. Figure 1 (A and B)
 148 shows an example of two unprocessed thermograms obtained from two independent cell
 149 cultures in the presence of three different concentrations of U(VI) nitrate.

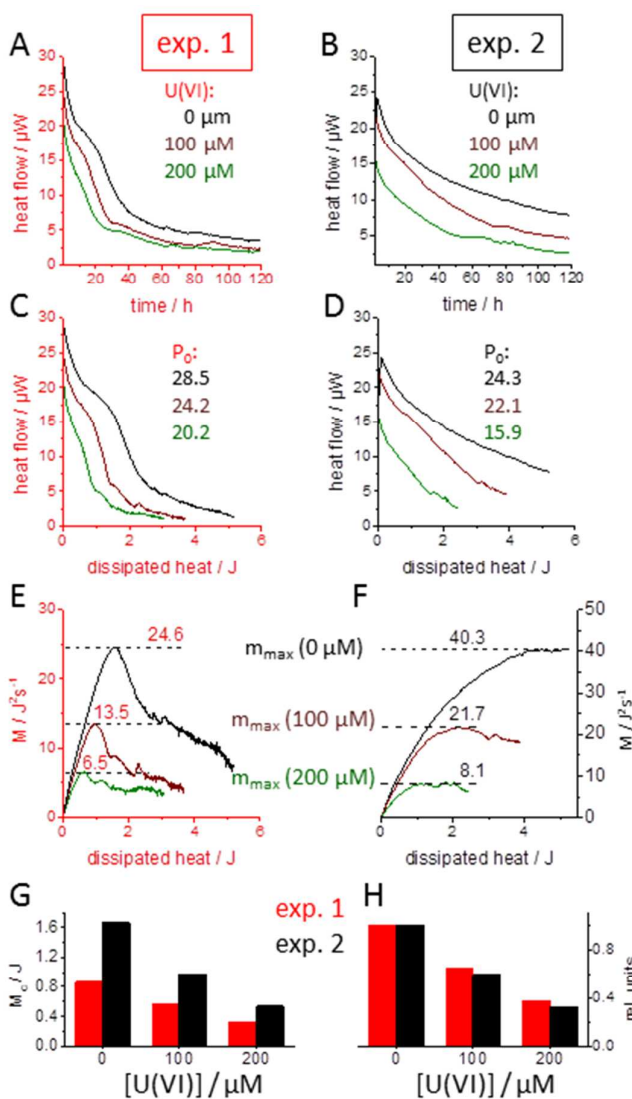


Figure 1: U(VI) dependence of the “metabolic capacity” for distinctly different temporal declines of metabolic activity. Thermograms from two independent *B. napus* cell cultures (left and right panels in A to F) were recorded at 0, 100 and 200 μM U(VI) (raw data in A and B), transformed into enthalpy plots (C and D) and the “metabolic capacity” M calculated (in J^2s^{-1} , E and F). G) The “characteristic metabolic capacity” M_c (in J) was obtained by dividing the M_{max} values by the initial thermal power P_0 of each thermogram (indicated in C and D). H) For comparison, the M_c values from panel G were scaled to unity with respect to the values obtained in the absence of U(VI).

152 All traces show a temporal decrease in metabolic thermal power and a general reduction
153 of heat production with increasing U(VI) concentration. Despite these robust general trends,
154 the shape of the metabolic decline was biphasic in one case (Fig. 1 A) and resembled an
155 exponential decay in the other (Fig. 1 B). Variability in the thermograms corresponding to
156 cells from different passages was generally observed but typically less than in the depicted
157 examples, where the standard deviation between thermograms (scaled to unity at time
158 zero) was 27%. Consequently, there is no obvious time point at which the comparison of
159 heat flow data would result in an unambiguous ranking of metabolic activity in dependence
160 of U(VI) concentration. Since enzyme activity and substrate supply are the salient
161 determinants of metabolic activity, we have searched for a quantitative descriptor related to
162 these two parameters in order to compare metabolic states.

163 Whereas enzyme activity generates thermal power (μW), the overall substrate depletion
164 is related to the total dissipated heat (J), because the latter is produced from substrate
165 consumption. We define the “metabolic capacity” M as the product of the time-dependent
166 heat flow P (overall “enzyme activity”) and the integrated dissipated heat H (overall
167 “substrate depletion”) relative to a time zero (t_0):

$$168 \quad M(t) := P(t) \cdot H(t), \quad \text{with} \quad H(t) = \int_{t_0}^t P(t) dt. \quad (1)$$

169 $M(H)$ can be calculated from an “enthalpy plot”, i.e., the plot of $\frac{dH}{dt}$ ($= P$) as a function of
170 H (Fig. 1 C and D), by multiplication of each original data pair. The result is graphically
171 displayed for the two independent experiments in Figs. 1 E and F. Since $H(t_0)$ and $P(t^\infty)$ are
172 zero, there will always be a characteristic time t_c and enthalpy $H_c = H(t_c)$ at which $M(H)$
173 reaches a maximum M_{\max} in units of J^2s^{-1} . Despite the differently shaped original
174 thermograms, the decrease of M_{\max} with increasing U(VI) toxicity was surprisingly similar for
175 the two experiments. Finally, the M_{\max} values were normalized with respect to the initial

176 heat flow $P_0 (= P(t_0))$ of each trace (Fig. 1 C and D), resulting in the “characteristic metabolic
177 capacity” M_c which carries the unit Joule (Fig. 1 G). The virtually identical U(VI) sensitivity of
178 the cells in the two experiments is best appreciated when the M_c values are further scaled to
179 those measured in the absence of uranium (Fig. 1 H). Whereas the raw data show the
180 metabolic activity at a given time point, the derivation of M_c contains the full history of the
181 thermogram up to the time t_c at which M_{max} was reached. The integrative evaluation of
182 thermograms according to Eq. 1 markedly reduced the variation in the M_c values between
183 the two experiments as compared to their original $P(t)$ values at any given time and U(VI)
184 concentration. Importantly, M_c is unambiguously defined and thus independent of
185 subjective choices of data points or physiological models for quantitating toxicity on the
186 basis of metabolic activity. The high reproducibility of the U(VI) sensitivity expressed by M_c
187 values derived from the differently shaped thermograms was surprising. In order to address
188 the physiological relevance of the purely phenomenologically derived value M_c , we analyzed
189 the calorimetric data from seven independent cell cultures (from which 37 thermograms
190 were evaluated) and asked whether the M_c values correlated with representative enzyme
191 activities which are typically used as markers for cell viability and are routinely measured by
192 an MTT test of oxidoreductase activity.¹⁴ The latter was performed either directly after the
193 calorimetric recordings, i. e., at low residual metabolic activity after M_c had been surpassed,
194 or in cell cultures outside the calorimeter under otherwise identical conditions. Calorimetric
195 and enzymatic data were averaged among the studied cultures.

196

197

198

199

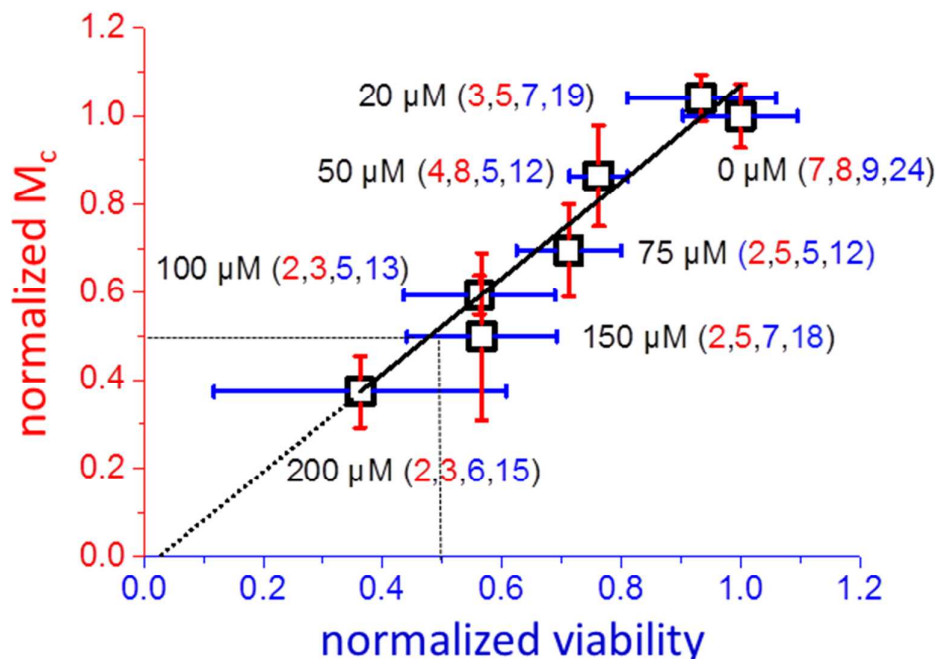


Figure 2: Correlation of oxidoreductase activity (viability) with “characteristic metabolic capacity” M_c . Data were averaged from experiments carried out with several cell cultures (Fig S3). The first and second number in brackets indicates how many cell cultures and thermograms contributed to each microcalorimetric data point, respectively. The third and fourth number give the corresponding information on cell cultures and individual MTT tests. For both assays, the data were normalized to the respective average of the values obtained with control samples in the absence of U(VI). Solid line: linear regression (slope of 1.09 / standard deviation 0.087; intercept -0.02 / standard deviation 0.065; Pearson’s $R = 0.98$). The average standard deviation in the calorimetric (red error bars) and oxidoreductase data (blue error bars) is 15.4 % and 21.9 %, respectively. Dotted lines: projection of half maximal inhibition on the regression line corresponds to an IC_{50} of $\sim 160 \mu\text{M}$ U(VI).

200 Figure 2 shows that oxidoreductase activity and M_c are to a very good approximation
 201 linearly related at U(VI) concentrations up to 200 μM . The evaluation clearly supports the
 202 physiological relevance of the definition of M_c . Therefore, we suggest the use of the
 203 “characteristic metabolic capacity” M_c as a calorimetric descriptor of metabolic activities
 204 under conditions, where metabolism does not follow simple growth models but declines
 205 continuously and with varying time dependence. In the case of *B. napus* cell culture, the M_c

206 values reveal a half-maximal inhibitory concentration of 160 μM U(VI) which agrees with the
207 MTT viability test due to the almost ideal correlation. The variance in the calorimetric data is
208 smaller and derived from integration over data interval of typically hours, as compared to
209 the MTT test which captures oxidoreductase activity at a subjectively chosen time point.
210 However, the data set is still too small to claim statistical significance (e.g. by a Students t-
211 test).

212 **Calculation of the U(VI) Speciation.** The effects of U(VI) on the cell metabolism may
213 depend on its speciation in the medium, i.e., its complexation with biomolecules, organic
214 and inorganic anions. The speciation is known to change with concentration, culture medium
215 composition, and pH value. Therefore, it was further analyzed under the conditions of the
216 calorimetric experiments (supporting information). Table 1 summarizes the main U(VI)
217 species in the medium R_{red} at pH 5.8 at different U(VI) concentrations (sample 1-6). Their pH
218 dependence is shown in Fig. S1.

219 Under the experimental conditions, the solid UO_2HPO_4 phase dominates at 20 μM $\text{U(VI)}_{\text{tot}}$
220 followed by a significant amount of $(\text{UO}_2)_3(\text{OH})_5^+$. The latter increases with $\text{U(VI)}_{\text{tot}}$ and is the
221 dominating species already at $\text{U(VI)}_{\text{tot}} > 50 \mu\text{M}$. Besides that, $(\text{UO}_2)_4(\text{OH})_7^+$, UO_2OH^+ , UO_2^{2+} ,
222 and $(\text{UO}_2)_2(\text{CO}_3)(\text{OH})_3^-$ species are formed, whereas the absolute amount of UO_2HPO_4 (s)
223 remains constant due to the limited phosphate concentration in medium R_{red} .

224 The calculations included only a single solid U(VI) phase. The significant formation of other
225 phases, i.e., metaschoepite or becquerelite, would be possible, however, can be neglected
226 under the applied conditions. All U(VI) solutions with medium R_{red} were treated by
227 ultracentrifugation (1 hours, 280,000 $\times g$). Due to the ultracentrifugation, a decrease of the
228 U(VI) concentrations in the solutions was observed (see Table S2 supporting information),
229 which can be attributed to the sedimentation of UO_2HPO_4 (s). This decrease is higher than

230 presumed assuming only the formation of $\text{UO}_2\text{HPO}_4(\text{s})$, however, lower than expected for
 231 the additional formation of significant amounts of metaschoepite and becquerelite (see
 232 Supporting Information). Thus, we assign the insoluble fraction of U(VI) mainly to
 233 $\text{UO}_2\text{HPO}_4(\text{s})$ and only small amounts of secondary mineral phases and / or U(VI) in colloids
 234 formed with nutrient components.

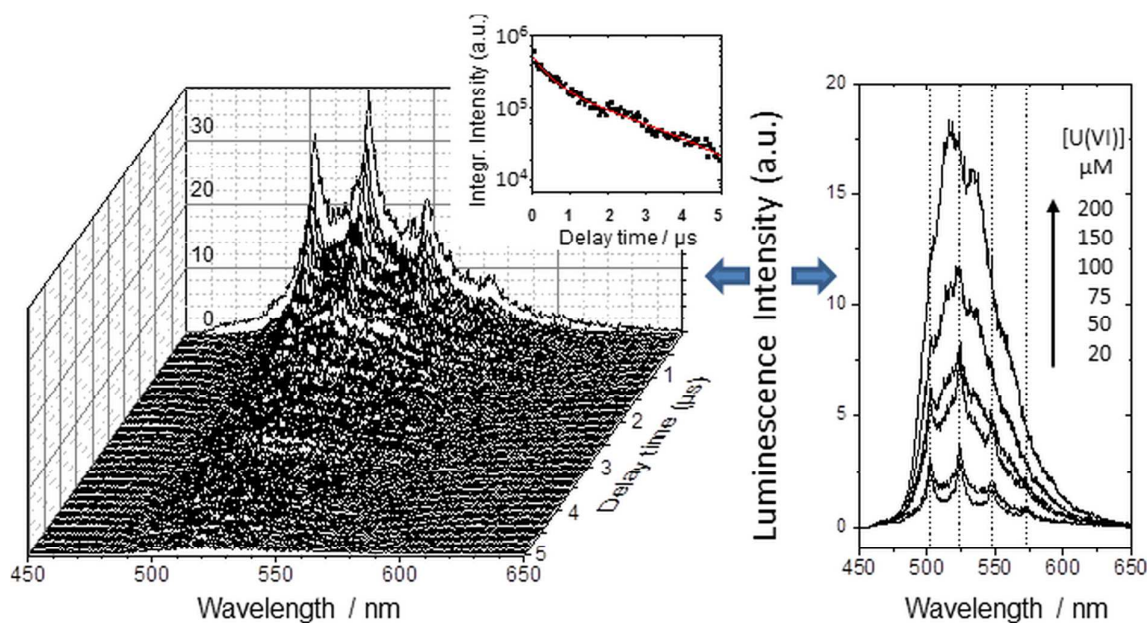
235 **Table 1:** Calculated U(VI) speciation in medium R_{red} with solid phases restricted to UO_2HPO_4
 236 (s) only (pH = 5.8, $p\text{CO}_2 = 10^{-3.5}$ atm). Species that do not exceed 4% of the total U(VI)
 237 concentration ($\text{U(VI)}_{\text{tot}}$) at any of the analyzed $\text{U(VI)}_{\text{tot}}$ values are not shown (see Supporting
 238 Information for further details).

Sample	$\text{U(VI)}_{\text{tot}}$ (mol/L)	$(\text{UO}_2)_2(\text{CO}_3)^-$ $(\text{OH})_3^-$	$(\text{UO}_2)_3(\text{OH})_5^+$	$(\text{UO}_2)_4(\text{OH})_7^+$	UO_2OH^+	UO_2^{2+}	$\text{UO}_2\text{HPO}_4(\text{s})$
1	2.0×10^{-5}	3.66	17.20	1.41	6.10	4.34	62.50
2	5.0×10^{-5}	5.42	49.00	7.70	4.70	3.34	25.00
3	7.5×10^{-5}	5.21	56.80	10.71	3.76	2.68	16.67
4	1.0×10^{-4}	4.94	60.50	12.80	3.18	2.26	12.50
5	1.5×10^{-4}	4.48	63.87	15.80	2.47	1.75	8.33
6	2.0×10^{-4}	4.13	65.50	17.90	2.05	1.46	6.25

239

240

241 **Experimental Analysis of U(VI) Speciation.** TRIFS measurements were carried out in
 242 medium R_{red} . Figure 3(A) shows the time-resolved luminescence spectrum at 50 μM U(VI)
 243 and the corresponding decay curve. The latter required bi-exponential approximation,
 244 indicating the presence of at least two luminescent species.



245

Figure 3: Time-resolved laser fluorescence spectroscopy. A) Spectral dependence of the time-resolved decay of the luminescence of 50 μM U(VI) in medium R_{red} (pH = 5.8) and the decay curve of the integral luminescence intensity (inset). B) Luminescence spectra of U(VI) in medium R_{red} as a function of the total U(VI) concentration (delay time: 51 ns). The vertical lines indicate the peak positions of the initial phosphate species.

246

Due to the presence of high amounts of organic substances as well as Fe³⁺ and Cl⁻ (cf.

247

Table S1, supporting information), dynamic quenching processes decreased the

248

luminescence lifetimes, thereby, hampering comparisons with model compounds.

249

Therefore, only the spectral shapes but not the lifetimes were further analyzed. Figure 3 (B)

250

shows U(VI) luminescence spectra recorded at a delay time of 51 ns at different

251

concentrations of U(VI)_{tot}. With increasing U(VI)_{tot}, the spectra changed significantly

252

indicating a change of the predominant U(VI) solution species. To attribute the spectra to

253

U(VI) species, spectra at different delay times were analyzed by peak deconvolution using

254

the peak fitting module of OriginPro 2015G and compared to literature data (Table S3,

255

supporting information).

256 The concentration-dependent U(VI) spectra measured after 51 ns (Fig. 3B) lead to specific
257 peak maxima obtained after spectral peak deconvolution. Whereas the peak maxima in
258 samples 1, 5, and 6 stayed constant with increasing delay times, the peaks in samples 2 and
259 3 showed a progressive hypsochromic shift over time, indicating the predominance of
260 different U(VI) species at different delay times (Figure S2, supporting information). The
261 occurrence of at least two luminescent solution U(VI) species was already concluded from
262 the rough lifetime estimation. The comparison of the peak maxima with those of reference
263 species (Table S3, supporting information) shows that U(VI) phosphate species dominated in
264 sample 1 and contributed also to the luminescence of sample 2-4 at 51 ns. At longer delay
265 times, the spectra corresponded to $(\text{UO}_2)_3(\text{OH})_5^+$ which was also the case for sample 5 and 6.
266 For the latter, the constancy of the spectra over all delay times demonstrated the
267 dominance of this species at $\text{U(VI)}_{\text{tot}} > 100 \mu\text{M}$. Although the luminescence lifetimes of all
268 samples were generally strongly quenched, a shorter lifetime of the U(VI) phosphate species
269 was observed than for the U(VI) hydroxo species (not shown). This agrees with literature
270 data, where luminescence lifetimes of $6.0 \mu\text{s}$ and $4.7 \pm 0.1 \mu\text{s}$ were reported for UO_2HPO_4
271 and $(\text{UO}_2)_x(\text{PO}_4)_y$, respectively¹⁶ in contrast to $19.8 \pm 1.8 \mu\text{s}$ for $(\text{UO}_2)_3(\text{OH})_5^+$.¹⁷ The TRIFS
272 results are further supported by the thermodynamic speciation calculations: UO_2HPO_4 and
273 $(\text{UO}_2)_3(\text{OH})_5^+$ dominates the U(VI) speciation at $20 \mu\text{M}$ and $\geq 50 \mu\text{M}$ $\text{U(VI)}_{\text{tot}}$, respectively (cf.
274 Table 1).

275 **U(VI) Speciation and Metabolic Activity.** Microcalorimetry has been used previously for
276 the assessment of U(VI)-dependent bacterial activities both in genetically engineered
277 bacteria¹¹ or communities found in uranium waste heaps.¹⁸ Typically, calorimetric studies
278 have been evaluated based on mathematically well-defined growth models¹⁹ or enthalpy
279 balances²⁰ in order to derive per cell estimates of metabolic rates. However, metabolic

280 calorimetry suffers from the lack of quantitative descriptors when simple growth models
281 cannot be applied. This is particularly the case with plant cells in the presence of heavy
282 metals which exhibit metabolic decline rather than growth. On the other hand, calorimetry is
283 particularly attractive for assessing radionuclide toxicity in the environmentally relevant low
284 dose exposure. Calorimetry is non-invasive, toxicity estimates can be derived directly from
285 the physical response of a living system, and experimental procedures enable minimal
286 handling efforts with radionuclides. In order to exploit the potential of microcalorimetry in
287 radioecological investigations, we have established a generally applicable evaluation of
288 thermograms from cells that exhibit metabolic decline without a preceding growth phase.
289 We have shown here that the “characteristic metabolic capacity” M_c (Eq. 1) is a reliable
290 calorimetric descriptor that scales with enzymatically determined viability in *B. napus* cells.
291 Using this descriptor, the influence of U(VI) speciation on metabolic activity can be
292 addressed. The TRLFS data and the thermodynamic modeling showed that the total amount
293 of U(VI) hydroxo complexes in fresh medium R_{red} increased with $U(VI)_{tot}$, whereas the low
294 phosphate concentration in medium R_{red} limited $UO_2HPO_4(s)$ formation to less than 13 μM .
295 Figure 4 correlates M_c with $U(VI)_{tot}$ and, alternatively, with the sum of the two hydroxo
296 complexes that make up more than 50% of all species for $[U(VI)_{tot}] > 20 \mu M$ (Table 1). The
297 normalized M_c values scale linearly with the U(VI) hydroxo complexes. Importantly, the data
298 intersect the y-axis nearly perfectly at the expected value of $M_c = 1$ for [U(VI) hydroxo
299 complexes] = 0. Although M_c varies linearly also with $[U(VI)_{tot}] > 0$, the required condition of
300 $M_c = 1$ for $[U(VI)_{tot}] = 0$ is not met. The data indicate that the dissolved hydroxo-species
301 possess highest bioavailability. In contrast, species that are not bioavailable ($UO_2HPO_4(s)$)
302 contribute to $U(VI)_{tot}$ without lowering M_c . This explains the right-shift of the plot of M_c vs.
303 $U(VI)_{tot}$. Correspondingly, metabolism is suppressed by 50% at a concentration of $\sim 90 \mu M$ of

304 the U(VI) hydroxo-complexes (intersection of regression line with x-axis, Fig. 4) which is
 305 significantly less than the ca. 160 μM estimated from the regression line in Fig. 2, where the
 306 indicated $[\text{U(VI)}]_{\text{tot}}$ concentration includes also non-bioavailable U(VI) species.

307 Speciation-dependence of U(VI) uptake and translocation is a general phenomenon. Both
 308 are affected by pH in *Arabidopsis thaliana* plants, although pH also altered the cellular redox
 309 balance in this organism.⁵ In contrast to the majority of invasive biochemical studies on U(VI)
 310 uptake and toxicity, isothermal microcalorimetry can provide a highly sensitive and non-
 311 invasive monitor of radionuclide-mediated metabolic effects. We have shown here for *B.*
 312 *napus* cells that metabolic responses of plant cells to U(VI) can be measured under the non-

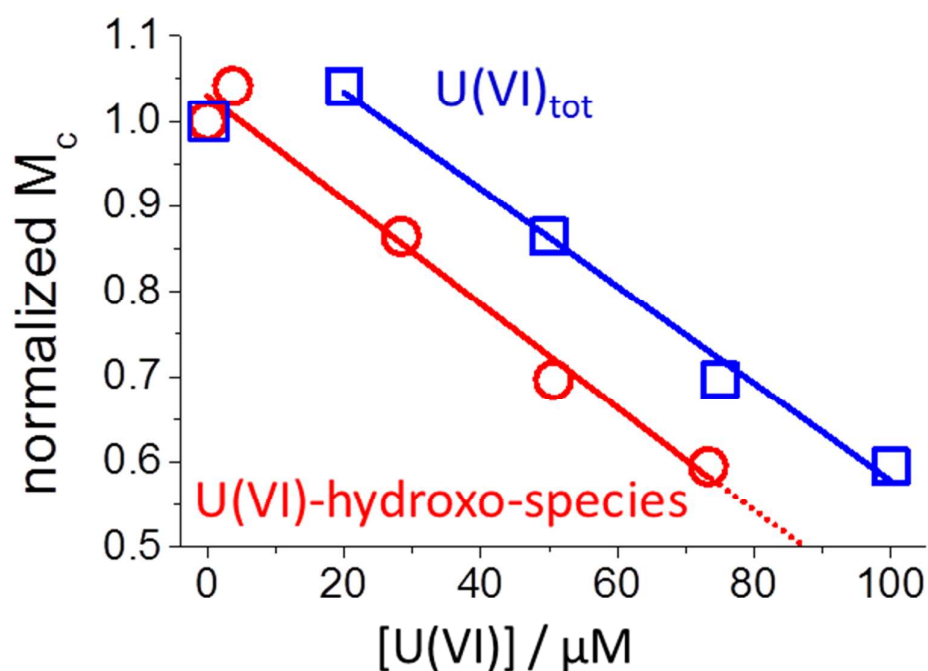


Figure 4: U(VI) species-dependency of the “characteristic metabolic capacity” M_c . Circles: the normalized M_c values (from Fig. 2) are plotted vs. the concentration of the dominant hydroxo-species, i.e., $[(\text{UO}_2)_3(\text{OH})_5^+] + [(\text{UO}_2)_4(\text{OH})_7^+]$. Squares: the same M_c values plotted vs. $\text{U(VI)}_{\text{tot}}$. Solid lines: linear regression of the data. The strong correlation (Pearson's $R = 0.99$) of M_c with the U(VI)-hydroxo-species (red) indicates that it is biologically the most relevant (dotted line: extrapolation to IC_{50} of $\sim 90 \mu\text{M}$).

313 photosynthetic regime of microcalorimetry. Remarkably, the “characteristic metabolic
314 capacity” M_c correlates sufficiently well with an established enzymatic viability test to be
315 used as a quantitative descriptor of toxicity. It appears less variable than the enzymatic test,
316 which is probably due to the integration of thermal data over time as opposed to the single
317 time point estimates obtained by the MTT test. The current observation is statistically not
318 significant (by Student’s t-test) but a systematic comparison of error margins in M_c
319 determinations and enzymatic assays will be interesting once larger data sets are available.
320 The current study demonstrates the potential of life cell microcalorimetry for radioecological
321 studies, enabling viability measurements independently of prior genetic or detailed
322 physiologic analyses. At the same time, handling of radionuclides can be reduced to an
323 absolute minimum. Although microcalorimetry has been shown to also reveal mechanistic
324 details on uranium toxicity when linked to genetic engineering¹¹, such data complement but
325 cannot replace studies on uranium-dependent specific enzyme activities (for example
326 Saenen et al. 2015). Instead, the strength of life cell microcalorimetry originates in its quick
327 and non-invasive systemic approach. It appears particularly suited for the future
328 quantification of toxicity mediated by internal exposure to other α - and β -emitting uranium
329 isotopes for which the present work provides the necessary reference. It remains to be
330 elucidated to which extent such investigations can contribute to the ecotoxicological risk
331 assessment of radionuclides, where endpoints at higher degrees of biological complexity
332 such as metabolism-driven processes (growth and development) may provide the missing
333 link to the effects of radionuclides at the molecular scale²¹.

334

335 **ACKNOWLEDGEMENTS**

336 The authors thank Dr. M. H. Obeid, J. Seibt, S. Heller, and J. Philipp for their valuable help
337 with the performance of the experiments, S. Gurlitt and B. Pfützner for ICP-MS
338 measurements, and S. Weiss for ultracentrifugation.

339

340 Associated Content

341 **SUPPORTING INFORMATION**342 **Thermodynamic Modeling of the U(VI) Speciation in Medium R_{red}**

343 Table S1, S2, Fig. S1

344 **Analysis of TRLFS Data**

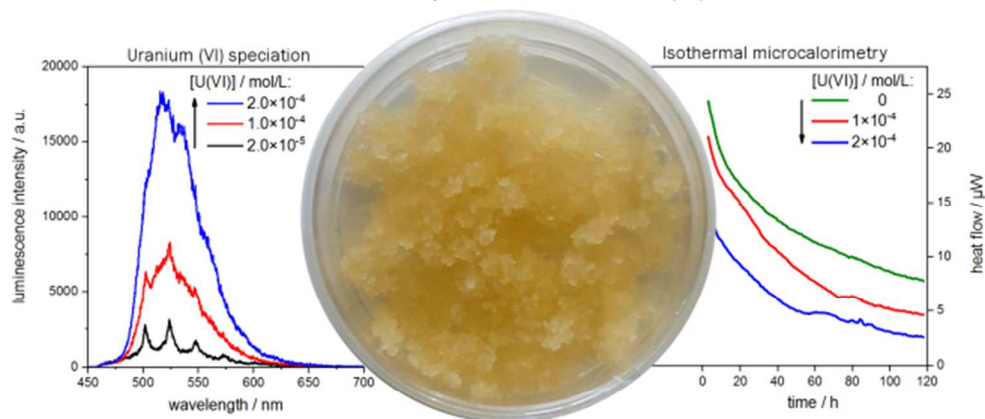
345 Table S3, Fig S2

346 **Dose response curves of metabolic capacity and oxidoreductase activity**

347 Fig. S3

348 **References**

- 349 1. Ebbs, S. D.; Brady, J.; Kochian, L. V. Role of uranium speciation in the uptake and
350 translocation of uranium by plants. *J. Exp. Bot.* **1998**, *49*, 1183-1190.
- 351 2. Laurette, J.; Larue, C.; Mariet, C.; Brisset, F.; Khodja, H.; Bourguignon, J.; Carrière, M.
352 Influence of uranium speciation on its accumulation and translocation in three plant species: Oilseed
353 rape, sunflower and wheat. *Environ. Exp. Bot.* **2012**, *77*, 96-107.
- 354 3. Laurette, J.; Larue, C.; Llorens, I.; Jaillard, D.; Jouneau, P.-H.; Bourguignon, J.; Carrière, M.
355 Speciation of uranium in plants upon root accumulation and root-to-shoot translocation: A XAS and
356 TEM study. *Environ. Exp. Bot.* **2012**, *77*, 87-95.
- 357 4. Misson, J.; Henner, P.; Morello, M.; Floriani, M.; Wu, T.-D.; Guerquin-Kern, J.-L.; Février, L.
358 Use of phosphate to avoid uranium toxicity in *Arabidopsis thaliana* leads to alterations of
359 morphological and physiological responses regulated by phosphate availability. *Environ. Exp. Bot.*
360 **2009**, *67*, 353-362.
- 361 5. Saenen, E.; Horemans, N.; Vanhoudt, N.; Vandenhove, H.; Biermans, G.; Van Hees, M.;
362 Wannijn, J.; Vangronsveld, J.; Cuypers, A. Effects of pH on uranium uptake and oxidative stress
363 responses induced in *Arabidopsis thaliana*. *Environ. Toxicol. Chem.* **2013**, *32*, 2125-2133.
- 364 6. Saenen, E.; Horemans, N.; Vanhoudt, N.; Vandenhove, H.; Biermans, G.; van Hees, M.;
365 Wannijn, J.; Vangronsveld, J.; Cuypers, A. Oxidative stress responses induced by uranium exposure at
366 low pH in leaves of *Arabidopsis thaliana* plants. *J. Environ. Radioact.* **2015**, *150*, 36-43.
- 367 7. Viehweger, K.; Geipel, G.; Bernhard, G. Impact of uranium(U) on the cellular glutathione
368 pool and resultant consequences for the redox status of U. *Biometals* **2011**, *24*, 1197-1204.
- 369 8. Günther, A.; Bernhard, G.; Geipel, G.; Reich, T.; Roßberg, A.; Nitsche, H. Uranium speciation in
370 plants. *Radiochim. Acta* **2003**, *91*, 319-328.
- 371 9. Geipel, G.; Viehweger, K. Speciation of uranium in compartments of living cells. *Biometals*
372 **2015**, *28*, 529-539.
- 373 10. Weiler, E.; Nover, L. *Allgemeine und molekulare Botanik*; Georg Thieme Verlag: Stuttgart,
374 2008; p 900.
- 375 11. Obeid, M. H.; Oertel, J.; Solioz, M.; Fahmy, K. Mechanism of attenuation of uranyl toxicity by
376 glutathione in *Lactococcus lactis*. *Appl. Environ. Microbiol.* **2016**, *82*, 3563-3571.
- 377 12. Zagorskina, N. V.; Goncharuk, E. A.; Alyavina, A. K. Effect of cadmium on the phenolic
378 compounds formation in the callus cultures derived from various organs of the tea plant. *Russ. J.*
379 *Plant Physiol.* **2007**, *54*, (2), 237-243.
- 380 13. Linsmaier, E. M.; Skoog, F. Organic Growth Factor Requirements of Tobacco Tissue Cultures.
381 *Physiol. Plantarum* **1965**, *18*, 100-127.
- 382 14. Mosmann, T. Rapid colorimetric assay for cellular growth and survival: application to
383 proliferation and cytotoxicity assays. *J. Immunol. Methods* **1983**, *65*, 55-63.
- 384 15. Lindl, T.; Gstraunthaler, G. *Zell- und Gewebekultur*; 6 ed.; Spektrum Akademischer Verlag:
385 Heidelberg, 2008.
- 386 16. Billard, I.; Geipel, G. Luminescence Analysis of Actinides: Instrumentation, Applications,
387 Quantification, Future Trends, and Quality Assurance. *Springer Ser. Fluoresc.* **2008**, *5*, 465-492.
- 388 17. Sachs, S.; Brendler, V.; Geipel, G. Uranium(VI) complexation by humic acid under neutral pH
389 conditions studied by laser-induced fluorescence spectroscopy. *Radiochim. Acta* **2007**, *95*, 103-110.
- 390 18. Schippers, A.; Hallmann, R.; Wentzien, S.; Sand, W. Microbial Diversity in Uranium-Mine
391 Waste Heaps. *Appl. Environ. Microbiol.* **1995**, *61*, 2930-2935.
- 392 19. Braissant, O.; Bonkat, G.; Wirz, D.; Bachmann, A. Microbial growth and isothermal
393 microcalorimetry: Growth models and their application to microcalorimetric data. *Thermochim. Acta*
394 **2013**, *555*, 64-71.
- 395 20. Maskow, T.; Paufler, S. What does calorimetry and thermodynamics of living cells tell us?
396 *Methods* **2015**, *76*, 3-10.
- 397 21. Fuller, N.; Lerebours, A.; Smith, J.T.; Ford, A.T. The biological effects of ionising radiation on
398 Crustaceans: A review. *Aquat. Toxicol.* **2015**, *167*, 55-67.
- 399

Brassica napus cells + uranium(VI)

TOC graphical abstract

84x47mm (200 x 200 DPI)