

Stability of U(VI) doped calcium silicate hydrate gel in repository-relevant brines studied by leaching experiments and spectroscopy

Originally published:

November 2018

Chemosphere 218(2019), 241-251

DOI: <https://doi.org/10.1016/j.chemosphere.2018.11.074>

Perma-Link to Publication Repository of HZDR:

<https://www.hzdr.de/publications/Publ-27649>

Release of the secondary publication
on the basis of the German Copyright Law § 38 Section 4.

CC BY-NC-ND

1 **Stability of U(VI) doped calcium silicate hydrate gel in repository-**
2 **relevant brines studied by leaching experiments and spectroscopy**

3 Jan-Martin Wolter^a

4 Katja Schmeide^{b,*}

5 Stephan Weiss^c

6 Frank Bok^d

7 Vinzenz Brendler^e

8 Thorsten Stumpf^f

9

10 All authors are from the Helmholtz-Zentrum Dresden - Rossendorf, Institute of Resource Ecology,

11 Bautzner Landstraße 400, 01328 Dresden, Germany

12

13 ^a j.wolter@hzdr.de

14 ^{b*} k.schmeide@hzdr.de, + 49 351 260 2436 corresponding author

15 ^c s.weiss@hzdr.de

16 ^d f.bok@hzdr.de

17 ^e v.brendler@hzdr.de

18 ^f t.stumpf@hzdr.de

19

20

21 Keywords: uranium, C-S-H, portlandite, carbonate, ionic strength, TRLFS

22

23 Colors should be used for print

24

25 **Abstract**

26 The stability of calcium silicate hydrate (C-S-H) gel doped with uranium to form calcium uranium
27 silicate hydrate (C-U-S-H) gel was investigated in 2.5 M NaCl, 2.5 M NaCl/0.02 M Na₂SO₄, 2.5 M
28 NaCl/0.02 M NaHCO₃ or 0.02 M NaHCO₃ solutions relevant to the geological disposal of radioactive
29 waste. The C-U-S-H gel samples were synthesized by direct U(VI) incorporation and characterized
30 with time-resolved laser-induced luminescence spectroscopy (TRLFS), infrared (IR) spectroscopy,
31 powder X-ray diffraction (XRD), scanning electron microscopy (SEM), differential scanning
32 calorimetry (DSC), and thermogravimetric analysis (TGA). Time-dependent pH changes as well as the
33 Ca, Si and U release from C-U-S-H gels into the brines, determined by inductively coupled plasma
34 mass spectrometry (ICP-MS), were monitored for three calcium-to-silicon (C/S) ratios (0.99, 1.55 and
35 2.02) over 32 d. Subsequently, changes of the U(VI) speciation and C-S-H mineralogy caused by
36 leaching were investigated with TRLFS, IR spectroscopy and XRD. Results indicated that
37 composition and pH value of the leaching solution, the presence of portlandite as well as formation
38 and solubility of calcite as secondary phase determine the U(VI) retention by C-S-H gel under high
39 saline and alkaline conditions. At high ionic strengths, the Ca release from C-S-H and secondary
40 phases like calcite is increased. Under hyperalkaline conditions only small amounts of U(VI) were
41 released during leaching. A decrease of the pH due to the additional presence of carbonate was linked
42 with an increased U(VI) release from C-S-H gel leading to the formation of aqueous calcium uranyl
43 carbonate in the supernatant solution.

44

45 **1. Introduction**

46 The long-term isolation of high-level nuclear waste, such as spent fuels from nuclear power
47 plants, in deep geological formations behind multiple protective barriers is recognized worldwide as
48 preferred strategy to protect humans and environment. In this concept, as part of the geo-engineered
49 barrier cementitious materials in the form of concrete or grout are foreseen to ensure mechanical
50 stability and sealing of disposal tunnels. Moreover, cementitious materials are commonly used for the
51 solidification of low and intermediate level radioactive waste. For a reliable long-term safety

52 assessment of a nuclear waste repository, it is necessary to identify materials and processes that
53 contribute to the retention of radionuclides potentially released due to water intrusion into a disposal
54 site. In addition, criteria that might lead to a mobilization of radionuclides have to be identified.

55 The immobilization potential of hardened cement paste (HCP) as well as of C-S-H, as main
56 component of HCP, towards radionuclides such as Cm(III), Th(IV) or U(VI) has been shown in a
57 number of literature studies, e.g. (Pointeau et al., 2004; Stumpf et al., 2004; Wieland et al., 1997).
58 Among the components of spent nuclear fuel 95.5% of the mass is contributed by the uranium isotopes
59 235, 236 and 238 (OECD, 2006). Their long half-lives and high chemo-toxicity render them important
60 long-term pollutants.

61 Studies on U-containing HCP showed the presence of uranium in the oxidation state +VI mainly
62 inside the C-S-H gel of HCP (Harfouche et al., 2006; Macé et al., 2013; Wersin et al., 2003; Zhao et
63 al., 2000). The structure of C-S-H gel is similar to a defected 14 Å tobermorite structure, consisting of
64 layered polyhedral CaO planes, silicate chains consisting of pairing and bridging Si tetrahedra called
65 “dreierketten” and interlayers filled with water and Ca²⁺ ions (Richardson, 2008). HCP and C-S-H
66 characteristics, such as layer-to-layer distance, length of silicate chains, and co-presence of portlandite
67 are defined by the molar C/S ratio of the cementitious material, which can exceed 1.7 in freshly
68 hydrated Portland cement and fall below 0.6 in nearly decomposed HCP (Grangeon et al., 2013;
69 Taylor, 1993; Thiery et al., 2013). The highest C/S ratio achieved by exclusively C-S-H in HCP is
70 around 1.7 while further Ca is contributed in the form of portlandite (Ca(OH)₂) (Yu et al., 1999).

71 Studies of C-U-S-H gel using TRLFS have identified 3 different U(VI) bonding environments: (i)
72 a surface complex, (ii) interlayer incorporation, (iii) a Ca uranate-like surface precipitate (Tits et al.
73 (2011; 2015)). An increase of the U(VI) retention by C-S-H gel with increasing C/S ratio was
74 associated with a U(VI) uptake mechanism that involves Ca from the C-S-H interlayers. Androniuk et
75 al. (2017) combined wet chemistry experiments and molecular dynamics computer simulations to
76 identify the sorption behavior of gluconate and U(VI) on C-S-H gel. The authors identified several
77 sorption sites for U(VI) on the C-S-H surface like mono and bidentate outer-sphere uranyl complexes
78 bound to deprotonated oxygens of bridging Si tetrahedra. Due to a sorption of Ca²⁺ ions on the same
79 sorption sites, a competition between uranyl and Ca²⁺ ions was suggested.

80 Gaona et al. (2012) developed a geochemical model for the interactions between U(VI) and C-S-
81 H gel based on the results of extended X-ray absorption fine structure (EXAFS) spectroscopy
82 (Harfouche et al., 2006) and TRLFS (Tits et al., 2011; Wieland et al., 2010). The authors confirmed
83 that U(VI) is mainly located in the C-S-H interlayers similar to Ca^{2+} ions. In contrast, the release of
84 U(VI) from HCP is controlled by dissolution and recrystallization of the C-U-S-H gel rather than by
85 surface desorption processes (Tits et al., 2015). It is therefore important to identify factors that will
86 affect the stability of C-U-S-H gel, such as pH, ionic strength and composition of the leaching
87 solution, the formation of secondary phases and how these might incorporate U(VI), as well as the
88 U(VI) species and their mobility in the leaching solution.

89 The majority of investigations on radionuclide retention by cementitious materials, however, have
90 only been performed at low ionic strengths, e.g. (Gaona et al., 2012; Harfouche et al., 2006; Tits et al.,
91 2011). Studies of cementitious materials in high ionic strength solutions include Hill et al. (2006),
92 Kienzler et al. (2010; 2016) and Bube et al. (2014). Hill et al. (2006) studied the dissolution of
93 radionuclide-free C-S-H gel (C/S 0.8–2.0) in 0.1 to 1 M NaCl solutions. The authors found an
94 increasing amount of leached Ca^{2+} with increasing NaCl concentration which was attributed to an ion
95 exchange mechanism between Na and Ca. Kienzler et al. (2010; 2016) and Bube et al. (2014) studied
96 the corrosion of U(VI)-containing cement monoliths in saturated NaCl- or MgCl_2 -rich brines.
97 Comprehensive analyses of drill core fragments and bulk powder samples after long-term leaching
98 showed the presence of a uranophane-like $(\text{Ca}(\text{UO}_2)_2(\text{SiO}_3\text{OH})_2 \cdot 5\text{H}_2\text{O})$ phase while no becquerelite,
99 meta-schoepite or di-uranate phases were detected. The U(VI) concentrations measured in leaching
100 solutions were up to 5×10^{-7} M while a diffusive penetration of halite into the cement monolith surface
101 up to 5 cm was observed.

102 Carbonate-containing high saline host rock pore waters, containing up to 4 M Na^+ , 2.8 M Cl^- ,
103 0.1 M SO_4^{2-} , and 0.06 M HCO_3^- , are reported for potential sites for nuclear waste repositories, e.g., for
104 Cretaceous argillites in North Germany (Peryt et al., 2010; Wolfgramm et al., 2011), as well as for
105 sedimentary bedrocks in Japan (Hama et al., 2007) and in Canada (Mazurek, 2004). However, to the
106 best of the author's knowledge, systematic studies on the stability of cementitious phases and their

107 retention potential for U(VI) in complex repository-relevant brines have not been reported in the
108 literature.

109 Such high ionic strengths pore waters could modify dissolution/recrystallization processes of C-S-
110 H gel including the formation of secondary phases, which in turn could affect the stability of C-S-H
111 gel as well as the U(VI) release or retention.

112 Thus, the objective of the present work was to evaluate the stability of C-U-S-H gel and its U(VI)
113 retention capability in high ionic strengths pore waters. For this, C-U-S-H gel samples with three
114 different C/S ratios (2.02, 1.55 and 0.99, representing a portlandite saturated C-S-H system as well as
115 chemically degraded cement paste) were prepared and leached in solutions containing 2.5 M NaCl,
116 0.02 M Na₂SO₄ and/or 0.02 M NaHCO₃. In addition, spectroscopic, microscopic and diffraction
117 techniques were applied before and after leaching to obtain a mechanistic understanding of the
118 underlying processes such as the evolution of the C-S-H gel and of U(VI) species on the solid phase as
119 well as secondary phase formation and formation of U(VI) species in the leaching solution.

120

121 **2. Materials and experimental methods**

122 **2.1. Materials**

123 The synthesis of C-U-S-H gel and subsequent leaching experiments were performed in an inert
124 gas glove box (N₂ atmosphere, O₂ < 2 ppm) using degassed deionized water (18 MΩ cm; mod. Milli-
125 RO/Milli-Q-System, Millipore, Schwalbach, Germany). Prior to experiments, centrifuge tubes
126 (polypropylene, Greiner bio-one, Kremsmünster, Austria) were immersed in 0.1 M HCl for 12 h and
127 subsequently rinsed with deionized water to remove contaminants.

128 Solid UO₃ was prepared from uranyl peroxide (Chemapol Ltd, Czech Republic) after a method
129 reported by Rosenheim and Daehr (1929). For C-S-H synthesis, KOH (ACS reagent, Roth, Karlsruhe,
130 Germany), NaOH (p. a., Roth), fumed silica (AEROSIL 300, Evonik, Essen, Germany), and
131 carbonate-free CaO (anhydrous, trace metals basis, Sigma-Aldrich, St. Louis, Missouri, USA) were
132 used. Previously, NaOH and KOH surface carbonate was removed by washing the NaOH and KOH
133 pellets several times with degassed deionized water in a Büchner funnel under inert gas atmosphere.

134 Total inorganic carbonate concentrations of prepared NaOH and KOH solutions were determined with
135 a multi-N/C 2100 S (Analytik Jena, Jena, Germany) and were below 100 μM . For drying, wet C-S-H
136 samples were either frozen in liquid nitrogen, and dried at 1 mbar and $-40\text{ }^\circ\text{C}$ for 1 d in a freeze-drying
137 system (mod. ALPHA 1-4 LSCplus, Christ, Osterode am Harz, Germany) or exposed to a CO_2 -free N_2
138 atmosphere over two weeks.

139

140 **2.2. Synthesis of C-U-S-H gel**

141 **2.2.1. Direct synthesis of C-U-S-H gel**

142 C-U-S-H gel samples (A–C) were prepared applying a direct incorporation method. For this, a
143 U(VI)-containing artificial cement pore water (ACW; (Berner, 1992)) was prepared by adding an
144 excess of solid UO_3 to a solution of 0.18 M KOH and 0.114 M NaOH (pH 13.3) and stirring for a
145 week. Subsequently, the ACW was separated from remaining solid UO_3 by centrifugation ($6800\times g$,
146 mod. Avanti J-20 XP, Beckman Coulter, Krefeld, Germany) and analyzed for U(VI) by ICP-MS
147 analysis (KED-mode, He gas, mod. NexION 350X/Elan 9000, PerkinElmer, Waltham, MA, USA)
148 after adjusting the pH to 2 with concentrated HNO_3 . The resulting ACW contained $19.5\text{ }\mu\text{M}$ U. This
149 U(VI)-containing ACW was added to a mixture of fumed silica and carbonate-free CaO. While the
150 CaO to SiO_2 ratio was varied depending on the targeted C/S ratio (1.0, 1.6, 2.0), the solid-to-liquid
151 (S/L) ratio was fixed at 24 g L^{-1} . Suspensions were shaken end-over-end with a rotator SB2 (Stuart,
152 Staffordshire, UK) for 14 d, and then filtered through a Whatman ashless 541-grade filter ($22\text{ }\mu\text{m}$ cut-
153 off). The C-U-S-H gel samples were washed with deionized water and stored moist in sealed tubes for
154 leaching experiments (cf. section 2.3) or dried for spectroscopy. For the determination of the elemental
155 compositions, C-S-H samples were decomposed in hydrofluoric acid and then analyzed for Ca, Si, Na,
156 K, and U by ICP-MS. An additional C-S-H gel (sample J) was synthesized at a S/L ratio of 24 g L^{-1} in
157 U(VI)-free ACW for SEM investigations that required a U(VI)-free C-S-H gel.

158

159

160 **2.2.2. Sorption of U(VI) onto C-S-H gel**

161 For a comparison of samples A–C with C-U-S-H gel synthesized by Tits et al. (2011), a further
162 C-U-S-H gel (CSH_{Nitrate}) with a target C/S ratio of 1.3 was prepared using the method proposed by
163 these authors. For this, sample J was exposed to a UO₂(NO₃)₂ solution which contained 19.5 μM
164 U(VI) for 14 d (S/L ratio: 24 g L⁻¹, pH after 14 d: 11.7). Subsequently, the C-U-S-H gel was similarly
165 filtered off, washed with deionized water, and stored moist.

166

167 **2.3. Leaching experiments**

168 For batch leaching experiments, wet C-U-S-H gel samples (A–C), prepared by direct synthesis,
169 were equilibrated in 2.5 M NaCl, 2.5 M NaCl/0.02 M Na₂SO₄, 2.5 M NaCl/0.02 M NaHCO₃ or 0.02 M
170 NaHCO₃ at a S/L ratio of 10 g L⁻¹ by shaking the samples end-over-end with a rotator SB2 for up to
171 32 d. Subsequently, the water contents of samples A–C were determined by TGA to determine the
172 actual S/L ratios in each leaching experiment. After different time intervals, up to 768 h, samples were
173 centrifuged at 6800×g. Each supernatant solution was analyzed for Ca, Si and U as well as for the pH
174 value. Solid and liquid phases after leaching were characterized by spectroscopic and further analytical
175 methods (cf. section 2.4.). All leaching experiments mentioned above were performed as triplicates.

176 Additionally, sample B was leached in 2.5 M NaCl/0.075 M NaHCO₃ for 768 h to visualize
177 changes in the TRLFS spectrum of C-U-S-H gel after leaching in solutions with high concentrations of
178 carbonate. To determine the influence of increasing carbonate concentrations on the pH evolution and
179 U(VI) release from C-U-S-H gel, sample A was exposed to solutions containing 1, 2, 10, 20, 50, and
180 75 mM NaHCO₃ for 768 h. Subsequently, the supernatants were separated by centrifugation and
181 analyzed for the pH value and released U. To get detailed information about the Ca and Si release in
182 dependence on NaCl concentrations, sample J was exposed to solutions containing 0, 0.5, 1.0, 1.5, 2.0,
183 and 2.5 M NaCl at an actual S/L ratio of 1.5 g L⁻¹ for 7 d. Subsequently, the supernatants were
184 analyzed for Ca and Si.

185

186 2.4. Analytical techniques and calculation methods

187 For TRLFS investigations, C-U-S-H gel samples were suspended in deionized water, transferred
188 into Boro 5.1 (ASTM type 1 class B glass, 5 mm) tubes (Deutero, Kastellaun, Germany) and shock-
189 frozen in liquid N₂. The tubes were placed in a cryogenic cooling system (mod. TG-KKK, KGW-
190 Isotherm, Karlsruhe, Germany) and further cooled with N₂ vapor to a constant temperature of
191 (153 ± 2) K. The cooling of the samples was necessary to generate a detectable U(VI) luminescence
192 signal by the suppression of dynamic quenching effects (Wang et al., 2005a). The laser set-up was
193 optimized for high intensities and good signal-to-noise ratios of the spectra: front entrance of 500 μm,
194 gate width of 2 ms and a delay time between 0.1 μs and 2 ms. To generate the excitation wavelength
195 of 266 nm, a Nd:YAG laser (mod. Minilite, Continuum, Santa Clara, CA, USA) with an average pulse
196 energy of 0.3 mJ was used. Emission wavelengths were detected between 14,947 cm⁻¹ (669 nm) and
197 27,100 cm⁻¹ (369 nm) by accumulating 100 laser pulses with a spectrograph (mod. iHR 550, HORIBA
198 Jobin Yvon, Unterhaching, Germany) and an ICCD camera (mod. ICCD-3000, HORIBA Jobin Yvon).
199 Positions of luminescence bands were determined by calculating the minima of the second derivative.

200 Immediately after removing the C-U-S-H gel from the N₂ atmosphere, IR samples were prepared
201 under ambient conditions by mixing approximately 1 mg of the dried samples with 300 mg dried KBr
202 and subsequently pressing for 2 min at 1×10⁹ Pa until clear pellets were obtained. IR samples were
203 measured with a Vertex 70/v spectrometer (Bruker, Billerica, Massachusetts, USA) equipped with a
204 D-LaTGS-detector (Lalanine doped triglycine sulfate), over a range of 400 to 4,000 cm⁻¹ in the
205 transmittance mode, with a spectral resolution of 4 cm⁻¹. Each spectrum was averaged over 64 scans.

206 Powder XRD data of C-U-S-H gel were collected with a MiniFlex 600 diffractometer (Rigaku,
207 Tokyo, Japan) equipped with a Cu Kα X-ray source (40 keV/15 mA) and the D/teX Ultra 1D silicon
208 strip detector. The diffractograms were recorded in the Bragg-Brentano θ -2 θ geometry at a scanning
209 speed of 0.6° per min. For sample preparation, the C-U-S-H gel was mounted as wet paste on a zero-
210 background Si sample holder with a Kapton window under inert gas conditions. The subsequent
211 Rietveld analysis of diffractograms was done with the program PDXL 2 (Rigaku) and the ICDD PDF-
212 4+ 2016 database.

213 For TGA and DSC investigations, the various C-S-H gels (sample J (freeze-dried), samples A–C
214 (wet pastes), sample C (dried under N₂ atmosphere)) were transferred in an Al₂O₃ crucible and
215 measured between 25 and 1000 °C at 20 K min⁻¹ in an Ar atmosphere with a STA 449 F5 Jupiter
216 thermal analysis instrument (Netzsch, Selb, Germany).

217 SEM imaging with secondary electrons was performed using a S-4800 microscope (Hitachi,
218 Tokyo, Japan) operated at an accelerating voltage of 10 kV. For sample preparation, the freeze-dried
219 C-S-H gel was put onto an aluminum stub covered with a carbon tab.

220 The pH values of solutions were measured with a pH meter (inoLab pH 720, WTW, Weilheim,
221 Germany) with a SenTix[®] Mic pH microelectrode (WTW) calibrated against standard pH buffers (pH:
222 6.865, 9.180 and 12.454) (WTW). The pH values of solutions with high ionic strengths were corrected
223 using a method reported by Altmaier et al. (2003; 2008).

224 The solubilities of calcite in 0–2.75 M NaCl solutions at pH 10, of portlandite in 0–3.2 M NaCl
225 solutions and of a C-S-H gel (C/S 1.1) in 0–3.2 M NaCl solutions at a S/L ratio of 1.5 g L⁻¹ were
226 calculated with the geochemical speciation code Geochemist's Workbench (GWB), Module React
227 (vers. 11.0.6) (Bethke, 2008) using the Pitzer ion interaction model. The thermodynamic database was
228 "THEREDA" (Thermodynamic Reference Database, <https://www.thereda.de>, Release No. 6). U(VI)
229 speciation in the carbonate-containing supernatant solution was also calculated with GWB using the
230 "PSI/Nagra Chemical Thermodynamic Database 12/07" (Thoenen et al., 2014). The parameter files
231 used for solubility and speciation calculations are added as supplementary material.

232

233 **3. Results and discussion**

234 **3.1. Characterization of C-U-S-H gel**

235 **3.1.1. Elemental composition**

236 C-U-S-H gel samples A–C, synthesized by the direct incorporation method, possessed C/S ratios
237 of 0.99, 1.55 and 2.02 (Table 1). From the Ca and Si amounts applied for C-S-H synthesis, ≤2.3%
238 remained in the ACW after the reaction time of 14 d (Table S1). That means, the C/S ratios after
239 syntheses are almost identical to the targeted C/S ratios. More than 96.5% of U, originally present in

240 the alkaline starting solution, was taken up by the C-S-H samples. U(VI) loadings were between
 241 1.34×10^{-4} and 2.29×10^{-4} mol kg⁻¹ which correspond to 31.9 to 54.5 ppm U(VI). The distribution
 242 ratios, $R_d = \frac{c_0 - c_{eq}}{c_{eq}} \cdot \frac{V}{m}$ [L kg⁻¹], increased from 1.2×10^3 to 1.2×10^4 L kg⁻¹ with increasing C/S ratio.

243 Compared to the distribution ratios reported by Tits et al. (2011) (1.0×10^3 to 1.0×10^6 L kg⁻¹),
 244 Poiteau et al. (2004) (3×10^4 to 1.5×10^5 L kg⁻¹) as well as determined for CSH_{Nitrate} (2.1×10^3 L kg⁻¹),
 245 the direct incorporation method provides comparable R_d values. While the sample CSH_{Nitrate} possesses
 246 a high U(VI) loading, its R_d value is similar to those of samples A–C. This is caused by a hydration
 247 effect and discussed in the supplementary material (Text S2).

248

249 **Table 1**
 250 Composition of C-U-S-H gel samples and distribution ratios (R_d) for U uptake by C-S-H gel.

Sample	C/S ratio	Ca [mol-%]	Si [mol-%]	Na [mol-%]	K [mol-%]	U [ppm]	R_d [L kg ⁻¹]
C	2.02	63.3	31.4	2.2	3.1	54.5	11700
B	1.55	53.6	34.5	4.9	6.8	37.3	4820
A	0.99	37.9	38.3	9.4	14.4	31.9	1169
J	1.35	53.7	39.9	2.9	3.6	<0.01	–
CSH _{Nitrate}	1.25	54.3	43.6	0.7	1.4	183	2130

251

252 **3.1.2. IR spectroscopic, powder XRD, TGA and DSC investigations**

253 Analyses of the samples A–C by IR (Fig. S2, Text S3), XRD (Fig. S3, Text S4), TGA and DSC
 254 (Fig. S4, Text S5) were entirely consistent with the formation of a tobermorite-like C-S-H gel with the
 255 co-existence of 18 mass-% portlandite for sample C (Text S1). Water contents of samples A–C were
 256 between 77 and 88% as shown in Table S2, Fig. S1 and Text S1.

257

258 **3.1.3. TRLFS investigations**

259 The TRLFS spectra and luminescence properties of the C-U-S-H gel samples are compiled in Fig.
 260 1 and Table S3, respectively. The first emission band in the TRLFS spectra of C-U-S-H is termed 0-0

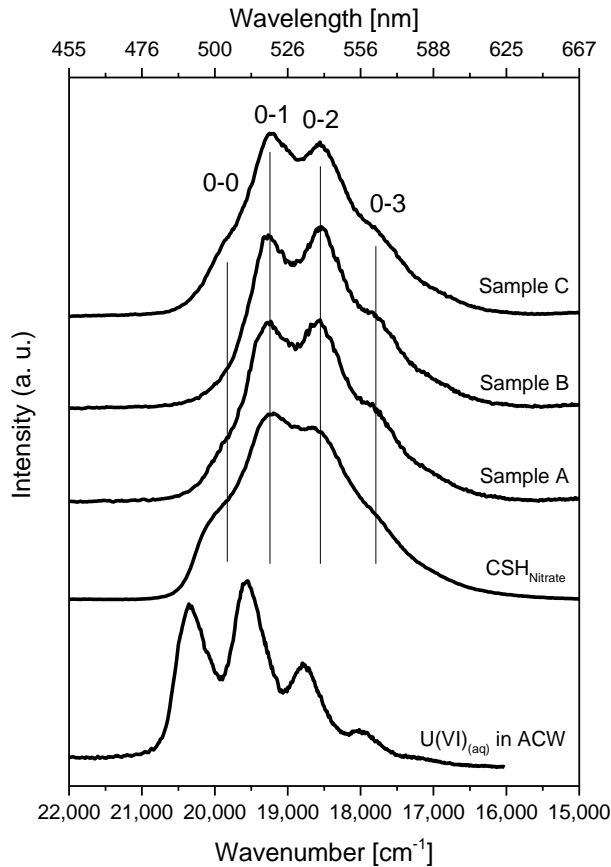
261 band and originates from the de-excitation of the $5f_{\delta}$ und $5f_{\phi}$ atomic orbital of U(VI) to the vibration
262 ground state of the σ_u molecular orbital, mainly dominated by axial bound O (Su et al., 2014). Further
263 de-excitations to vibration coupled states of σ_u are termed 0-1, 0-2 and so on.

264 A shift of emission bands towards lower energies (red-shift) was reported in the literature for an
265 increase of electron density of the uranyl ligands (Liu and Beitz, 2006; Tits et al., 2011). This can be
266 caused by a ligand change from e.g. water against hydroxide ions in ACW or a sorption of U(VI) on
267 the C-S-H surface whereby hydroxide ligands are substituted against silicate. An incorporation of
268 uranyl in the C-S-H interlayer should even further red-shift the emission bands due to the complete
269 absence of hydroxide ligands and the exclusive presence of silicate ligands in the C-S-H interlayer.
270 The electron density from relevant ligands increases in the following order: $\text{H}_2\text{O} < \text{OH}^- < \text{SiO}_4^{4-}$.

271 Observed luminescence emission bands of U(VI) dissolved in ACW (Fig. 1) are in good
272 agreement with bands reported for the uranyl hydroxide complex, $\text{UO}_2(\text{OH})_4^{2-}$, occurring in ACW
273 under comparable conditions (Tits et al., 2011) (Table S3). The 0-0 band of sample C is located at
274 $19,992 \text{ cm}^{-1}$, thus between the 0-0 bands of the U(VI)/C-S-H surface complex ($20,150 \text{ cm}^{-1}$) and the
275 U(VI)/C-S-H incorporated species ($19,844 \text{ cm}^{-1}$), reported by Tits et al. (2011) for a C-U-S-H gel with
276 a C/S ratio of 1.07. Compared to the further emission lines of samples A–C, the 0-0 band is weakly
277 pronounced. Since the 0-0 band is located in higher proximity to the 0-0 band of U(VI) coordinated by
278 hydroxide in ACW, it is concluded that the 0-0 band in samples A–C and $\text{CSH}_{\text{Nitrate}}$ is mainly caused
279 by a U(VI) species located on the C-S-H surface where OH^- and SiO_4^{4-} ligands are present.

280

281



282

283 **Fig. 1.** TRLFS spectra of samples C, B, A and CSH_{Nitrate} (C/S: 2.02, 1.55, 0.99 and 1.25) before leaching as well as U(VI)_(aq)
 284 (19.5 μM) in ACW.
 285

286 Note that the 0-0 band of the reference sample CSH_{Nitrate}, synthesized according to the procedure
 287 reported by Tits et al. (2011), is more distinct than in samples A–C. Due to the surface exposure of
 288 sample J to uranyl nitrate during the synthesis of CSH_{Nitrate}, U(VI) is probably first sorbed on the C-S-
 289 H surface before it is incorporated in the C-S-H interlayer by recrystallization. This was also described
 290 by Tits et al. (2011) who observed an ongoing incorporation of U(VI) in the C-S-H crystal structure
 291 over a time period of 6 months after U(VI) surface sorption on C-S-H gel. Due to the shorter synthesis
 292 time of 14 d compared to 6 months, CSH_{Nitrate} probably possesses a higher amount of surface sorbed
 293 U(VI) which is also higher compared to samples A–C where U(VI) is mainly located in the C-S-H
 294 interlayers.

295 The positions of the 0-1 and 0-2 bands of samples A–C are distinct around 19,300 and
 296 18,500 cm⁻¹, respectively, and most probably belong to uranyl incorporated in the C-S-H interlayer.
 297 Thus, during the 14 d of synthesis of the C-U-S-H gel the direct incorporation method ensures a fast

298 incorporation of U(VI) in the C-S-H gel whereas the U(VI) sorption procedure requires longer time
299 periods (up to 6 months) to acquire a high ratio of U(VI) incorporated in the C-S-H interlayer. Thus,
300 for the preparation of C-U-S-H gel, representative for cement encapsulated nuclear waste, within short
301 times (14 d) the direct incorporation method is preferable.

302 Luminescence decay lifetime analyses of samples A–C show the presence of two U(VI) species,
303 one with short lifetimes of 103–119 μs and one with long lifetimes of 521–604 μs (Table S3). These
304 lifetimes are in the same range as the lifetimes determined for the U(VI)/C-S-H surface complex
305 ((205–293) \pm 50 μs) and the U(VI)/C-S-H incorporated species ((546–647) \pm 50 μs) reported by Tits
306 et al. (2011). However, it should be noted that the lifetime of U(VI) species in solid C-S-H gel is
307 influenced by sample preparation, C/S ratio, moisture of the sample, and presence of quenchers. Thus,
308 a species identification based on the lifetime is only reasonable to a limited extent (Chisholm-Brause
309 et al., 2001; 2004).

310 From the TRFLS study it can be concluded that the U(VI) in samples A–C is predominantly
311 incorporated into the interlayer structure of C-S-H gel and to a much smaller degree sorbed on the C-
312 S-H gel surface.

313

314 **3.2. Leaching of C-U-S-H gel**

315 The results of batch leaching experiments, between 18 and 768 h exposure time, are shown for
316 sample B (C/S 1.55) in terms of Ca, Si and U release into the supernatant solution (Fig. 2[a], Fig. S5,
317 Fig. S6). The actual S/L ratios during leaching, that differed from the applied S/L ratio of 10 g L⁻¹ due
318 to the bulk water content of the wet gels, were determined with 1.2, 1.5 and 2.3 g L⁻¹ for samples A, B
319 and C, respectively (Table S2).

320 For all samples (A–C), final Ca, Si and U concentrations as well as pH values of the supernatant
321 solutions after 768 h leaching are given in Table 2. Between 18 and 768 h leaching, the pH values
322 varied with \pm 0.1 around the values given in Table 2. For the calculation of the percentage dissolution
323 of samples A–C, the Ca, Si and U concentrations for a hypothetical complete dissolution at actual S/L
324 ratios are given in Table S4.

325

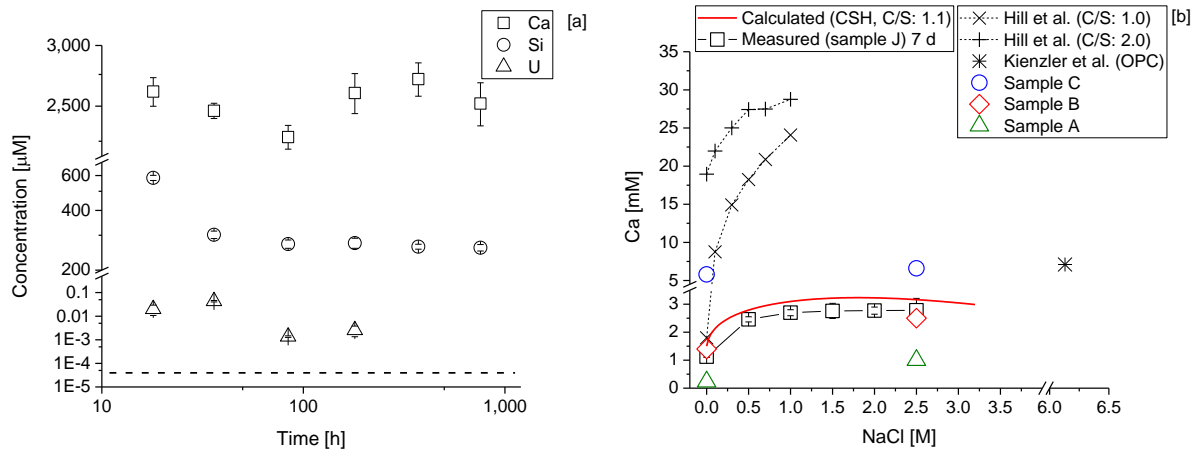
326 **3.2.1. Leaching in 2.5 M NaCl and 2.5 M NaCl/0.02 M Na₂SO₄**

327 The contact of samples A–C with a 2.5 M NaCl leaching solution leads to a Ca concentration in
328 solution of about 1.0–6.6 mM (Fig. 2, Table 2). Consequently, the pH value of the leaching solution
329 increased from 7 to 11.7–12.2.

330 Hill et al. (2006), who monitored the Ca release from radionuclide-free C-S-H gel into 0 to 1 M
331 NaCl solutions over 1 month (Table S5 ([Ca]: 2–29 mM leached, C/S: 1.0–2.0, S/L: 10–77 g L⁻¹),
332 showed that the presence of 1 M NaCl increased the amount of Ca leached from C-S-H gel compared
333 to pure water. The authors suggested an ion exchange mechanism between Na and Ca which increased
334 the Ca release from C-S-H gel in the presence of 1 M NaCl. They also found a good agreement of the
335 Ca release from portlandite and C-S-H gel with a high C/S ratio in 1 M NaCl.

336 Compared to this study, the results of the present study show a lower Ca mobilization into
337 solution which is attributed to the lower S/L ratios (1.2 to 2.3 g L⁻¹) and salting in/salting out effects
338 that have an impact on the Ca solubility. To study the C-S-H gel solubility behavior at low S/L ratios,
339 sample J (C/S 1.35) was exposed to salt solutions between 0 and 2.5 M NaCl at a S/L ratio of 1.5 g L⁻¹
340 over 7 d. Between 0 and 1.5 M NaCl, an increase of the Ca release from 1.1 mM towards 2.7 mM
341 (Fig. 2[b]) was observed. At higher NaCl concentrations, the Ca release stagnated around 2.7 mM. The
342 Si release of sample J reached a maximum of 0.95 mM between 0.5 and 1 M NaCl (Fig. S7). Leaching
343 results of samples A–C in pure water were in good agreement with these observations and showed a
344 lower release of Ca, especially for C-S-H gel with a low C/S ratio, compared to the leaching in 2.5 M
345 NaCl (Fig. 2[b]). Additionally, the Ca release from a C-S-H gel with a C/S ratio of 1.1 was calculated
346 at a S/L ratio of 1.5 g L⁻¹ for 0 to 3.2 M NaCl solutions (Fig. 2[b]). The calculations showed an
347 increase of the Ca release from the C-S-H gel up to a maximum at 1.5 M NaCl and a slight decrease at
348 higher NaCl concentrations. This is in good agreement with the leaching results of samples B and J.
349 Therefore, in terms of Ca release from C-S-H gel the effect of a 2.5 M NaCl solution is comparable to
350 a 1 M NaCl solution.

351



352

353 **Fig. 2.** Calcium, silicon and uranium concentration in the supernatant solution after leaching of sample B in 2.5 M NaCl as a
 354 function of time (ICP-MS detection limit for U is $4.2 \times 10^{-5} \mu\text{M}$ (dashed line)) [a]. Calculated solubility of a C-S-H
 355 gel with a C/S ratio of 1.1 in NaCl solutions between 0 and 3.2 M at a S/L ratio of 1.5 g L^{-1} [b] (calculation cf.
 356 section 2.4). Measured calcium release from sample J (S/L 1.5 g L^{-1}) between 0 and 2.5 M NaCl and from samples C
 357 (S/L 2.3 g L^{-1}), B (S/L 1.5 g L^{-1}) and A (S/L 1.2 g L^{-1}) in water and 2.5 M NaCl in comparison to literature data of
 358 Hill et al. (2006) (S/L 50 g L^{-1}) and Kienzler et al. (2010) (OPC = ordinary Portland cement) [b].
 359

360 The Si release of sample B into the 2.5 M NaCl solution amounts to 2.1% of total Si (0.26 mM)
 361 after 768 h (Fig. 2[a]). Please note that due to the release of Ca and Si from the C-S-H gel, the final
 362 C/S ratios of the C-S-H gel samples after leaching are decreased in dependency on the initial C/S
 363 ratios and the composition of the leaching solutions (Table 2). To ease the recognition of the various
 364 samples, however, we continue to use the C/S ratios before leaching in the following text. Generally,
 365 an increased Ca and Si release from concrete at higher ionic strength is equivalent to a faster chemical
 366 degradation of HCP in a nuclear waste repository.

367

368 **Table 2**
 369 Calcium, silicon and uranium concentrations and pH values of supernatant solutions as well as C/S ratios determined after
 370 768 h leaching of sample C (C/S 2.02, S/L 2.3 g L⁻¹), sample B (C/S 1.55, S/L 1.5 g L⁻¹) and sample A (C/S 0.99, S/L
 371 1.2 g L⁻¹). Values in brackets give the percentage dissolution of the ions present in the respective samples.

	Leaching solutions				
	Water	2.5 M NaCl	2.5 M NaCl/ 0.02 M Na ₂ SO ₄	2.5 M NaCl/ 0.02 M NaHCO ₃	0.02 M NaHCO ₃
Ca [mM]					
Sample C	5.8 (13.0)	6.6 (14.8)	7.4 (16.6)	0.21 (0.5)	0.034 (0.1)
Sample B	1.4 (7.5)	2.5 (13.4)	2.4 (12.8)	0.16 (0.9)	0.062 (0.3)
Sample A	0.23 (2.4)	1.0 (10.3)	1.9 (19.6)	0.18 (1.9)	0.17 (1.8)
Si [mM]					
Sample C	0.029 (0.1)	0.11 (0.5)	0.068 (0.3)	3.8 (17.1)	5.3 (23.9)
Sample B	0.13 (1.1)	0.26 (2.1)	0.26 (2.1)	2.4 (19.8)	8.4 (69.4)
Sample A	0.48 (4.9)	0.64 (6.5)	0.48 (4.9)	2.6 (26.5)	8.6 (87.8)
U [μM]					
Sample C	7.8×10 ⁻⁴	< 4.2×10 ⁻⁵	< 4.2×10 ⁻⁵	0.16 (4.6)	0.022 (0.6)
Sample B	8.7×10 ⁻⁴	< 4.2×10 ⁻⁵	< 4.2×10 ⁻⁵	0.20 (12.5)	0.64 (40.0)
Sample A	1.1×10 ⁻³	< 4.2×10 ⁻⁵	< 4.2×10 ⁻⁵	0.18(16.4)	0.24 (21.8)
pH					
Sample C	12.0	12.2	12.0	10.4	10.3
Sample B	11.7	11.9	11.7	10.1	9.6
Sample A	11.4	11.7	11.5	9.9	9.3
C/S ratio after leaching					
Sample C	1.75	1.57	1.69	–	–
Sample B	1.45	1.37	1.38	–	–
Sample A	1.01	0.96	0.84	–	–

372

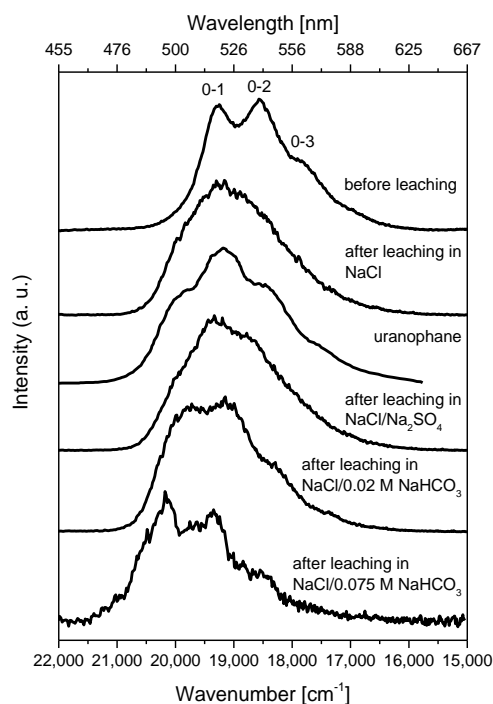
373 Only 2.7% (4.3×10^{-2} μM) of the total U(VI) of sample B was released in the first 36 h of leaching
 374 (Fig. 2[a]). After 186 h, the U(VI) concentration decreased below the ICP-MS detection limit of
 375 4.2×10^{-5} μM, indicating a reincorporation of U(VI) into the C-S-H gel or into a secondary phase
 376 formed.

377 Kienzler et al. (2016) investigated the U(VI) release from U(VI)-containing cementitious material
378 prisms submerged in highly saline salt solutions (5.98 M NaCl, 0.02 M CaSO₄, 0.02 M MgSO₄) over
379 32 years and observed an increase of uranium concentrations from $2 \times 10^{-3} \mu\text{M}$ towards $5 \times 10^{-1} \mu\text{M}$
380 over that time period. Despite the differences in solution composition, mineral assemblages and S/L
381 ratios, both experiments showed only small amounts of mobilized uranium due to the presence of
382 saline solutions.

383 To clarify the U(VI) speciation for the leached solids in the present study, TRLFS was applied
384 (Fig. 3). The spectrum of sample B leached in 2.5 M NaCl is broadened and shifted towards higher
385 energies indicating the presence of a uranophane-like phase with a maximum around $19,230 \text{ cm}^{-1}$ (Fig.
386 3). The presence of a diuranate-type U(VI) mineral would shift the maximum of the spectrum towards
387 lower energies around $18,100 \text{ cm}^{-1}$ (Tits et al., 2011) and was therefore excluded.

388 These findings correspond well with the observations of Kienzler et al. (2010) who characterized
389 concrete drill dust after leaching by XRD, TRLFS and EXAFS and reported the presence of a
390 uranophane-like phase.

391



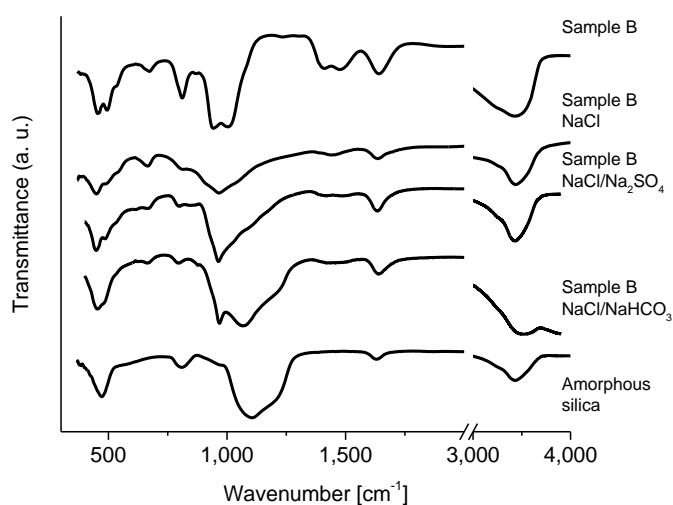
392

393 **Fig. 3.** TRLFS spectra of sample B (C/S 1.55) before and after 768 h of contact with different leaching solutions (2.5 M
394 NaCl, 2.5 M NaCl/0.02 M Na₂SO₄, 2.5 M NaCl/0.02 M NaHCO₃, 2.5 M NaCl/0.075 M NaHCO₃) in comparison to
395 the spectrum of uranophane (Kienzler et al., 2010).

396

397 The IR spectrum of sample B leached in 2.5 M NaCl shows a decrease of band resolution due to a
398 reduction of the C/S ratio and crystallinity due to leaching (Fig. 4). The Si–OH vibration at 960 cm^{-1}
399 was not shifted which indicates that the silanol (Si-OH) groups with bridging tetrahedra of the C-S-H
400 structure remained intact after leaching in 2.5 M NaCl.

401



402

403 **Fig. 4.** IR spectra of sample B (C/S 1.55) before and after 768 h of contact with different leaching solutions (2.5 M NaCl,
404 2.5 M NaCl/0.02 M Na₂SO₄, 2.5 M NaCl/0.02 M NaHCO₃) in comparison to the spectrum of amorphous silica
405 (AEROSIL 300).
406

407 In summary, due to increased ionic strength (2.5 M NaCl) the Ca and Si release from C-S-H gel
408 increased compared to pure water while the U(VI) environment partially changed from C-S-H towards
409 a uranophane-like phase. Nonetheless, the structure of the C-S-H gel remained mostly stable and
410 almost no U(VI) was released into the aqueous phase.

411 With regard to safety assessment of HCP barriers in deep geological waste disposals for actinides,
412 after contact with about 2.5 M NaCl solutions, a uranophane-like phase becomes the mobility
413 controlling phase for U(VI) on the HCP surface.

414 The Ca, Si and U release of sample B in 2.5 M NaCl/0.02 M Na₂SO₄ after 768 h (Fig. S5) was
415 similar to the Ca, Si and U release in 2.5 M NaCl (Table 2). Moreover, TRLFS and IR investigations
416 of C-U-S-H gel leached in 2.5 M NaCl/0.02 M Na₂SO₄ show no further changes compared to the
417 leaching in 2.5 M NaCl (Fig. 3, Fig. 4). Only the Si release kinetic in the first 18 h of leaching is

418 decreased due to the additional presence of 0.02 M Na₂SO₄ (Fig. 2[a], Fig. S5). This could be caused
419 by the adsorption of sulfate on the C-S-H surface as reported by (Divet and Randriambololona, 1998).
420 After 768 h of leaching, the Si concentrations in both leaching experiments matched. Conclusively,
421 0.02 M Na₂SO₄ does not have any negative effects on C-S-H stability or U(VI) retention by C-S-H gel.

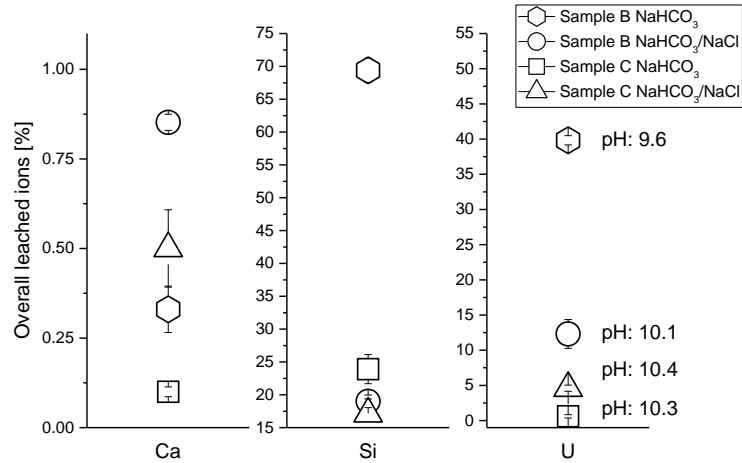
422

423 **3.2.2. Leaching in 2.5 M NaCl/0.02 M NaHCO₃ and 0.02 M NaHCO₃**

424 The presence of carbonate increased the amount of leached Si for samples A–C to 17.1-26.5%
425 (2.4–3.8 mM) in 2.5 M NaCl/0.02 M NaHCO₃ and to 23.9-87.8% (5.3–8.6 mM) in 0.02 M NaHCO₃
426 (Fig. S6, Table 2, Fig. 5). Thus, for all samples a partial destabilization of the silicate chains of the C-
427 S-H structure leading to the formation of SiO_{2(am)} can be assumed. This was confirmed by IR
428 spectroscopy due to the appearance of a shoulder at 1,234 cm⁻¹ which is also present in the IR
429 spectrum of SiO_{2(am)} (Fig. 4). For sample C leached in carbonate-containing solutions, the SiO_{2(am)}
430 band was only weakly pronounced (Fig. S8), suggesting that a part of the NaHCO₃ preferentially
431 reacted with portlandite, consequently, less SiO_{2(am)} was formed. However, the IR spectrum of sample
432 C leached in 2.5 M NaCl/0.02 M NaHCO₃ shows a distinct carbonate band around 1,400 cm⁻¹ (Fig.
433 S8) that probably belongs to calcite (Sato and Matsuda, 1969) formed by reaction of portlandite and
434 carbonate. In contrast, only a weak carbonate band around 1,400 cm⁻¹ is observed in the IR spectrum
435 of sample B leached in 2.5 M NaCl/0.02 M NaHCO₃, indicating a low amount of solid CaCO₃. This
436 implies that, compared to Ca(OH)₂, the C-S-H gel reacts to a lesser extent with carbonate.

437 Compared to carbonate-free leaching experiments, the leaching of sample B in 2.5 M
438 NaCl/0.02 M NaHCO₃ causes a lower Ca concentration (0.9%, 0.16 mM) as well as a lower pH (10.1)
439 in the leaching solution (Table 2, Fig. 5). Under carbonate-free conditions, the pH in a concrete system
440 is buffered by the dissolution of portlandite. In the presence of HCO₃⁻, the precipitation of CaCO₃
441 removes Ca²⁺ and OH⁻ from the solution, thus, lowering the pH and Ca concentrations compared to
442 carbonate-free solutions. Final C/S ratios of leached samples were not determined since the Ca content
443 of C-U-S-H gel after leaching is distributed between C-S-H and CaCO₃.

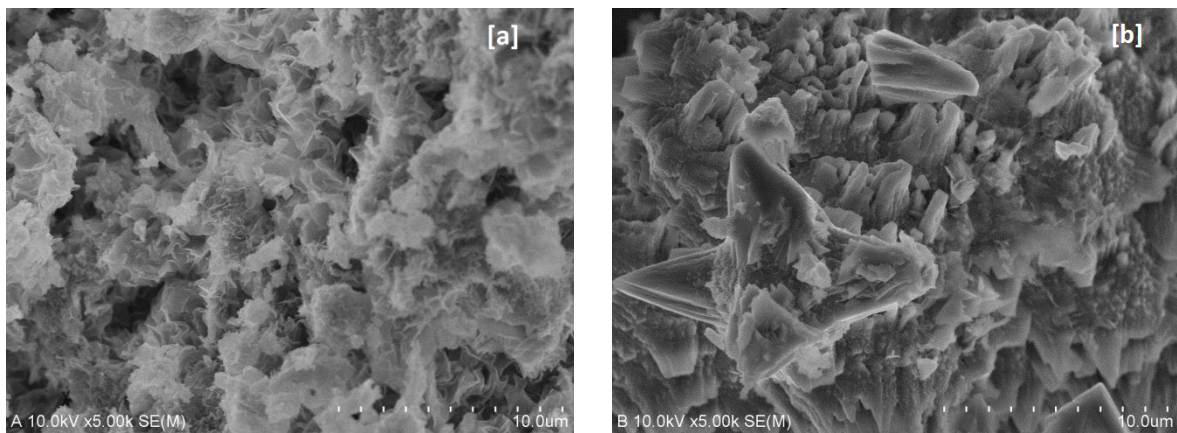
444



445

446 **Fig. 5.** Percentage amount of leached calcium, silicon and uranium with corresponding pH values after leaching of sample B
 447 (C/S 1.55) and sample C (C/S 2.02) in 0.02 M NaHCO₃ or 2.5 M NaCl/0.02 M NaHCO₃ for 768 h.
 448

449 Before leaching, the SEM image shows the typical C-S-H fiber structure (Ashraf and Olek, 2016)
 450 (Fig. 6[a]). After leaching in 0.02 M NaHCO₃, the overgrowth of the C-S-H surface of sample J with
 451 euhedral calcite crystals (Ashraf and Olek, 2016) was observed (Fig. 6[b]).



452

453 **Fig. 6.** SEM images of sample J (C/S 1.35) before leaching [a] and after leaching in 0.02 M NaHCO₃ [b].
 454

455 The U(VI) release of sample B due to leaching in 2.5 M NaCl/0.02 M NaHCO₃ is strongly
 456 increased to 12.5% (0.2 μM, pH 10.1) of the overall U(VI) (circle in Fig. 5, Table 2). This effect is
 457 even more distinct in the absence of 2.5 M NaCl where the 0.02 M NaHCO₃ leaching solution
 458 mobilizes 40% of the U(VI) of sample B (hexagon in Fig. 5, Table 2). In contrast, sample C shows a
 459 lower amount of released U(VI) of 4.6% (0.16 μM, pH 10.4) in 2.5 M NaCl/0.02 M NaHCO₃ and

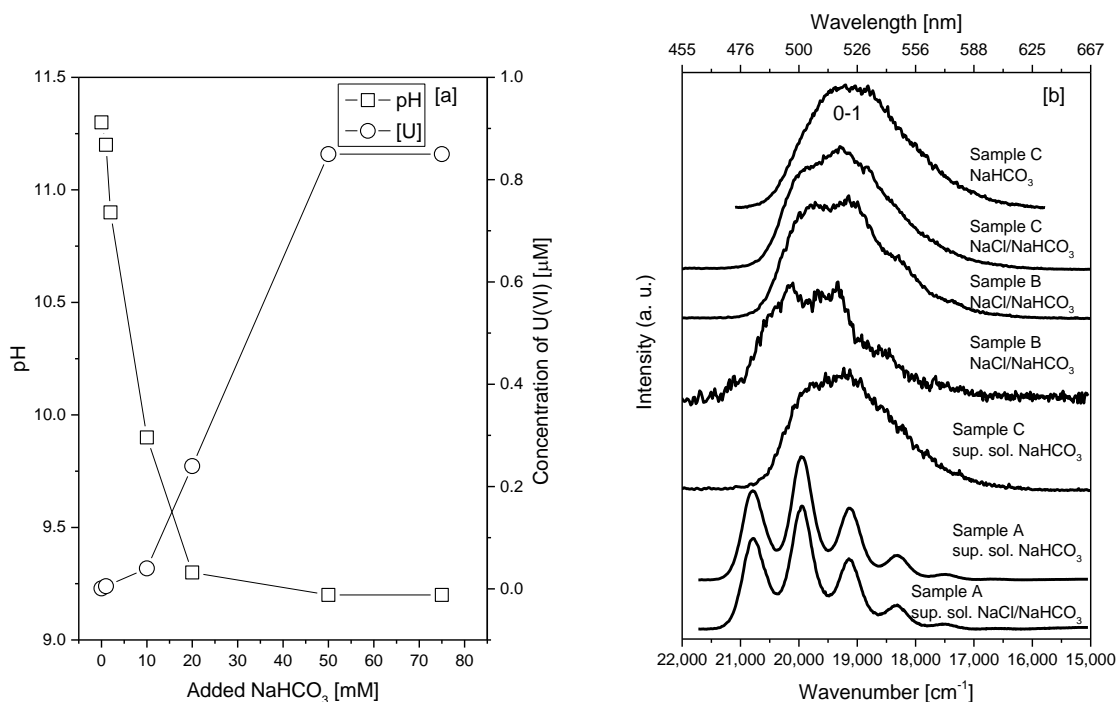
460 0.6% in 0.02 M NaHCO₃ (triangle/square in Fig. 5, Table 2). These results are a clear indication for an
461 increased U(VI) mobility in the presence of carbonate and a reduction of the pH which influences the
462 U(VI) speciation in the leaching solution.

463 Under hyperalkaline (pH > 12) conditions, U(VI) hydroxide complexes such as UO₂(OH)₄²⁻ or
464 UO₂(OH)₃⁻ dominate U(VI) speciation and solubility (Torrents, 2014), whereas a lower pH (8 < pH <
465 10.6) soluble uranyl carbonate complexes (e.g., UO₂(CO₃)₃⁴⁻) can occur in carbonate-containing
466 aqueous systems (Wang et al., 2004). In the presence of 0.034-0.21 mM Ca, determined in solution in
467 the current study, ternary calcium uranyl carbonate complexes, Ca₂UO₂(CO₃)₃(aq) and CaUO₂(CO₃)₃²⁻,
468 would likely form (Bernhard et al., 2001). These ternary species impede the U(VI) sorption onto solids
469 as shown, for instance, for ferrihydrite and quartz (Fox et al., 2006), clay (Joseph et al., 2013) and
470 granite (Schmeide et al., 2014).

471 To investigate U(VI) speciation changes in the leaching solution from UO₂(OH)₄²⁻ or UO₂(OH)₃⁻
472 towards aqueous calcium uranyl tricarbonatate in dependence on pH and amount of NaHCO₃, a leaching
473 series of sample A in solutions containing up to 75 mM NaHCO₃ was performed (Fig. 7[a]). The
474 leaching solutions of sample A show a non-linear decrease of the pH value coupled with increasing
475 amounts of released U in dependence on the NaHCO₃ contents. TRLFS investigations confirmed the
476 predominance of aqueous calcium uranyl tricarbonatate (Fig. 7[b]) for NaHCO₃ concentrations ≥
477 75 mM which corresponds to pH 9.3. This complex possessed well-defined emission bands at 20,803,
478 19,964, 19,157 and 18,348 cm⁻¹ and a lifetime of 750 μs (at 153 K) similar to the data reported by
479 Wang et al. (2004) (20,812, 19,952, 19,131 and 18,315 cm⁻¹; 1282 μs (at 6 K)) (Table S2). This
480 complex was also observed in the presence of 2.5 M NaCl when the pH and carbonate content were
481 similar (Fig. 7[b]).

482 Speciation calculations performed for the composition determined in the supernatant solution of
483 sample B after leaching in 2.5 M NaCl/0.02 M NaHCO₃ ([Ca]: 0.159 mM, [Si]: 2.4 mM, [U]: 0.2 μM,
484 [carbonate]: 14 mM, pH: 10.1), supported the formation of aqueous calcium uranyl tricarbonatate
485 complexes.

486 At higher pH values, the high concentration of OH⁻ suppresses the formation of calcium uranyl
 487 tricarbonate, and spectra with low intensities and without any band resolution are detected (sample C
 488 sup. sol. NaHCO₃, Fig. 7[b]).
 489

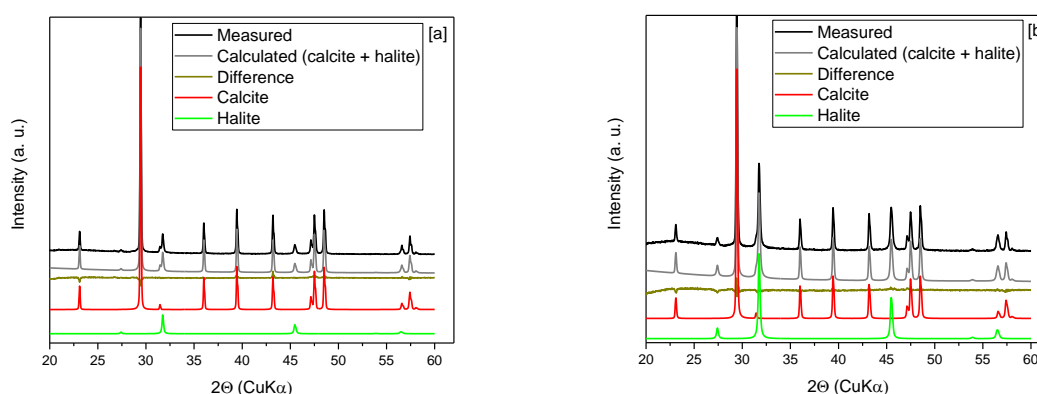


490
 491 **Fig. 7.** Evolution of pH value and U concentration of the supernatant solution of sample A as function of added NaHCO₃,
 492 (768 h leaching, S/L 1.2 g L⁻¹) [a]. TRLFS spectra of sample C leached in 0.02 M NaHCO₃, sample C leached in
 493 2.5 M NaCl/0.02 M NaHCO₃, sample B leached in 2.5 M NaCl/0.02 M NaHCO₃, sample B leached in 2.5 M
 494 NaCl/0.075 M NaHCO₃, supernatant solution of sample C leached in 0.02 M NaHCO₃, supernatant solution of
 495 sample A leached in 0.1 M NaHCO₃, supernatant solution of sample A leached in 2.5 M NaCl/0.075 M NaHCO₃,
 496 (768 h leaching) [b].
 497

498 Nonetheless, further factors such as formation of secondary CaCO₃ phases like calcite, vaterite
 499 and aragonite and their solubility in dependence on ionic strength may also influence the U(VI)
 500 mobility due to sorption of U(VI) on these CaCO₃ polymorphs. To investigate the influence of the
 501 secondary phase formation, TRLFS spectra of leached samples A–C were recorded. The TRLFS
 502 spectra of sample B leached in 2.5 M NaCl/0.02 M NaHCO₃ or 2.5 M NaCl/0.075 M NaHCO₃ (Fig.
 503 7[b]) show a shift of the 0-1 band towards higher energies from 19,318 cm⁻¹ to 19,980 cm⁻¹ and
 504 20,234 cm⁻¹, respectively. These shifts towards higher energies can be explained by an increased
 505 amount of U(VI) sorbed or incorporated in calcite formed as secondary phase as reported in the
 506 literature (Elzinga et al., 2004; Geipel et al., 1997; Smith et al., 2015).

507 TRLFS investigations of a natural U(VI)-containing calcite showed 0-0 and 0-1 bands at 20,790
508 and 20,000 cm^{-1} , respectively (Wang et al., 2005b). Since the leached samples contain a mixture of
509 U(VI) sorbed on calcite and U(VI) remaining in the C-S-H gel, the recorded spectra after leaching in
510 carbonate-containing solutions are caused by an overlap of both U(VI) coordination environments.

511 Rietveld analyses of the XRD patterns of sample C and A leached in 2.5 M NaCl/0.02 M
512 NaHCO_3 (Fig. 8) show, that C-U-S-H gel and portlandite are partly converted to CaCO_3 whereby
513 exclusively calcite was detected.



514
515 **Fig. 8.** Powder XRD patterns of samples C (C/S 2.02) [a] and A (C/S 0.99) [b] after leaching in 2.5 M NaCl/0.02 M
516 NaHCO_3 , (768 h leaching), calcite (database card number 01-083-4601), halite (database card number 00-005-0628),
517 (Cu $K\alpha$ X-ray source).
518

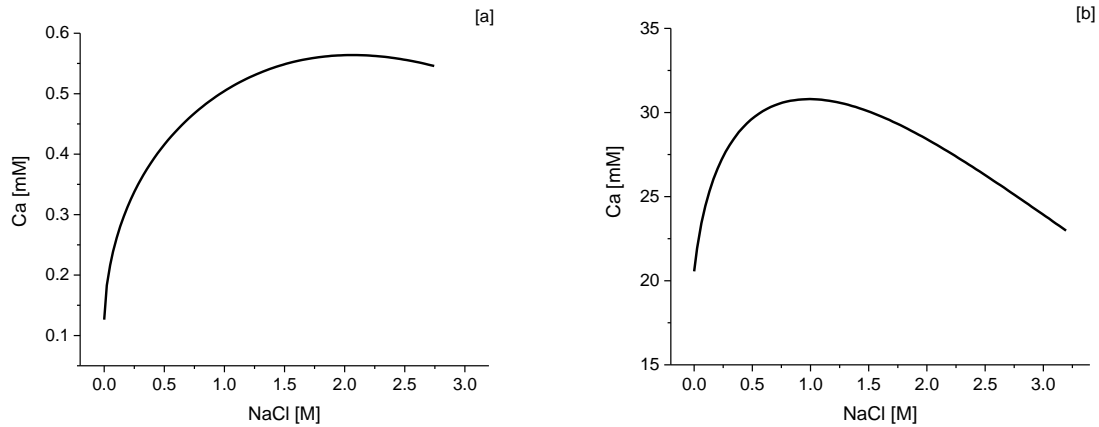
519 XRD patterns of samples C and A leached in 0.02 M NaHCO_3 but absence of NaCl (Fig. S9) also
520 suggest that the samples are partly converted into secondary CaCO_3 phases: A part of sample A is
521 converted into 70% calcite and 30% metastable aragonite while sample C, that consists of C-U-S-H
522 gel and portlandite, is partly converted into 80% calcite and 20% metastable vaterite (Table S6). A
523 comparison with data from (Black et al., 2007), who investigated carbonation effects in
524 mechanochemically prepared C-S-H gels with C/S ratios between 0.2 and 1.5 under ambient
525 conditions, revealed the occurrence of predominantly vaterite for samples with C/S ratios ≥ 0.67 while
526 in C-S-H samples with C/S ratios ≤ 0.5 mainly aragonite was detected. Although this trend is similar to
527 that of the present study, the absence of calcite and the lower C/S boundary for the vaterite formation
528 confirm differences between both studies. This can probably be attributed to different preparation
529 methods and carbonation techniques of both studies. Further studies where similar cementitious
530 systems were exposed to carbonate detected either the simultaneous presence of calcite, vaterite and

531 aragonite (Chang and Fang, 2015; Ibanez et al., 2007) or a reversed trend where aragonite was
532 observed at higher and vaterite at lower C/S ratios (Auroy et al. (2018)). These findings suggest that
533 the formation of the various CaCO_3 modifications depends on the cement system and the carbonation
534 method. The results of the present study also suggest that high amounts of NaCl suppress the
535 formation of metastable CaCO_3 modifications such as vaterite and aragonite. This is tentatively
536 attributed to an increased recrystallization rate of these phases as suggested by Takita et al. (2007).

537 TRLFS investigations of sample C leached in exclusively 0.02 M NaHCO_3 showed a broad
538 spectrum without any band resolution that is less blue-shifted than the spectrum of sample B leached
539 in 0.02 M NaHCO_3 (Fig. 7[b]). Due to the higher amount of portlandite in sample C, combined with
540 the higher solubility of portlandite compared to C-S-H gel, most of the carbonate is precipitated as
541 calcite and vaterite (Fig. S9). Thus, the amount of carbonate in solution, which could contribute to a
542 decomposition of the C-U-S-H structure, is reduced. Simultaneously, a higher pH value in solution
543 after CaCO_3 precipitation is maintained. Since portlandite does not contain U(VI) its dissolution does
544 not contribute to a U(VI) mobilization. Calcite, however, acts as a U(VI) sink as implied by the
545 TRLFS investigations (Fig. 7[b]). Compared to metastable CaCO_3 phases like vaterite, better U(VI)
546 retention properties were reported for calcite (Noubactep et al., 2006).

547 Compared to the leaching of sample C in 0.02 M NaHCO_3 solution, the additional presence of
548 2.5 M NaCl increased the amounts of released U(VI) and Ca from 0.6 to 4.6% and from 0.1 to 0.5%,
549 respectively (square/triangle, Fig. 5), although the pH values were with 10.3 and 10.4 very similar.
550 Solubility calculations (cf. section 2.4) showed that the calcite solubility increases with increasing
551 NaCl concentrations between 0 and 2.5 M NaCl (Fig. 9[a]). Also the solubility of portlandite increases
552 between 0 and 1 M NaCl, but decreases afterwards (Fig. 9[b]). If, in turn, less calcite precipitates,
553 which also leads to somewhat higher Ca^{2+} and CO_3^{2-} concentrations in the supernatant, less U(VI) can
554 be immobilized on the calcite surface. Thus, the formation of aqueous calcium uranyl tricarbonate is
555 favored against the formation of U(VI) hydroxide complexes at a pH around 10.4.

556



557

558 **Fig. 9.** Calculated solubilities of calcite in 0–2.75 M NaCl solutions at pH 10 [a] and of portlandite in 0–3.2 M NaCl
 559 solutions [b]. Details of calculations are given in section 2.4.

560

561

562 4. Conclusions

563 The stability of C-U-S-H gel and its U(VI) retention capability in high ionic strengths pore waters
 564 was studied. For this, C-U-S-H gel samples with three different C/S ratios (0.99, 1.55 and 2.02) were
 565 synthesized applying a direct incorporation method which ensured an U(VI) incorporation
 566 predominantly into the interlayer structure of C-S-H.

567 These C-U-S-H gel samples were leached in solutions containing 2.5 M NaCl, 0.02 M Na₂SO₄
 568 and/or 0.02 M NaHCO₃, thus, reflecting conditions expected for host rock pore waters of potential
 569 nuclear waste repositories. The results showed that the composition of the leaching solution had a
 570 direct influence on the alteration of the C-S-H structure and evolution of the pH value in these
 571 systems.

572 Compared to pure water, the leaching in a 2.5 M NaCl solution increased the release of Ca from
 573 C-S-H gel from 1.4 to 2.5 mM (initial C/S ratio: 1.55, S/L ratio: 1.5 g L⁻¹). The effect of a 2.5 M NaCl
 574 solution was comparable to a 1 M NaCl solution. Despite the release of Ca, IR investigations
 575 suggested that the C-U-S-H structure remained stable, while U(VI) retention was largely unaffected
 576 due to a formed uranophane-like phase detected by TRLFS. The additional presence of 0.02 M
 577 Na₂SO₄ did not show any effect on the stability of the C-U-S-H structure.

578 In the presence of carbonate (0.02 M), the U(VI) retention was coupled to the alteration stage of
579 the C-S-H structure as well as to the pH evolution of the leaching solution. Carbonate preferentially
580 reacted with portlandite to form calcite and vaterite while the C-U-S-H structure remained intact,
581 which was confirmed by XRD and IR investigations. Furthermore, the absence of a distinct blue-shift
582 of the U(VI) luminescence spectrum of the leached portlandite-rich sample confirmed the persistent
583 presence of U(VI) in the C-U-S-H gel. In the case of C-U-S-H gel with a lower C/S ratio, representing
584 altered concrete, carbonate removed Ca from solution which subsequently led to a further release of
585 Ca from the C-U-S-H gel. Thus, the polyhedral CaO plane of the C-U-S-H structure was increasingly
586 destabilized, as verified by the appearance of Si-O bands of amorphous silica in the IR spectrum. The
587 thereby formed secondary CaCO₃ phases, detected by XRD after leaching, contributed to a certain
588 extent to U(VI) retention, as shown by a blue-shift of the U(VI) luminescence spectrum of the leached
589 solid. Nevertheless, the destabilization of the polyhedral CaO plane of the C-U-S-H structure led to an
590 enhanced release of U(VI). Due to an overall lower pH value in the presence of carbonate, aqueous
591 calcium uranyl tricarbonate was formed as confirmed by TRLFS.

592 The U(VI) release from the portlandite-rich C-U-S-H sample into carbonate-containing solutions
593 was slightly higher in the presence than in the absence of NaCl. This was attributed to an increased
594 calcite solubility at enhanced salinities, which led to higher Ca²⁺ and CO₃²⁻ concentrations in solution,
595 and thus, to formation of aqueous calcium uranyl tricarbonate.

596 In the context of a nuclear waste repository, the high immobilization potential of C-U-S-H gel for
597 U(VI) appears to be unaffected by highly saline groundwaters. In the presence of carbonate, however,
598 it is expected that the release of U(VI) from C-U-S-H gel will be greatly enhanced.

599

600

601 **Acknowledgements**

602 The German Federal Ministry for Economic Affairs and Energy (BMWi) is thanked for financial
603 support (no. 02 E 11415B). The authors would like to thank Dr. Harald Foerstendorf (HZDR), Dr.
604 Gerhard Geipel (HZDR), Dr. Atsushi Ikeda-Ohno (HZDR), Dr. Jan Tits (PSI) and Dr. Barbara
605 Lothenbach (Empa) for the helpful discussions and advice as well as K. Heim, S. Beutner, B. Pfützner

606 and A. Scholz (all from HZDR) for IR, ICP-MS and SEM measurements. Furthermore, the authors
607 would like to specifically thank the two anonymous reviewers for their valuable comments.

608

609 **Appendix A. Supplementary material**

610 Supplementary data associated with this article can be found, in the online version, at

611 **References**

612 Altmaier, M., Metz, V., Neck, V., Müller, R., Fanghänel, T., 2003. Solid-liquid equilibria of
613 $\text{Mg}(\text{OH})_2(\text{cr})$ and $\text{Mg}_2(\text{OH})_3\text{Cl}\cdot 4\text{H}_2\text{O}(\text{cr})$ in the system Mg-Na-H-OH-O-Cl- H_2O at 25 °C. *Geochim*
614 *Cosmochim Acta* 67, 3595-3601.

615 Altmaier, M., Neck, V., Fanghänel, T., 2008. Solubility of Zr(IV), Th(IV) and Pu(IV) hydrous oxides
616 in CaCl_2 solutions and the formation of ternary Ca-M(IV)-OH complexes. *Radiochim Acta* 96, 541-
617 550.

618 Androniuk, I., Landesman, C., Henocq, P., Kalinichev, A., 2017. Adsorption of gluconate and uranyl
619 on C-S-H phases: Combination of wet chemistry experiments and molecular dynamics simulations for
620 the binary systems. *Phys Chem Earth* 99, 194-203.

621 Ashraf, W., Olek, J., 2016. Carbonation behavior of hydraulic and non-hydraulic calcium silicates:
622 potential of utilizing low-lime calcium silicates in cement-based materials. *J Mater Sci* 51, 6173-6191.

623 Auroy, M., Poyet, S., Le Bescop, P., Torrenti, J.-M., Charpentier, T., 2018. Comparison between
624 natural and accelerated carbonation (3% CO_2): Impact on mineralogy, microstructure, water retention
625 and cracking. *Cement Concrete Res* 109, 64-80.

626 Berner, U.R., 1992. Evolution of pore water chemistry during degradation of cement in a radioactive
627 waste repository environment. *Waste Manage* 12, 201-219.

628 Bernhard, G., Geipel, G., Reich, T., Brendler, V., Amayri, S., Nitsche, H., 2001. Uranyl(VI) carbonate
629 complex formation: Validation of the $\text{Ca}_2\text{UO}_2(\text{CO}_3)_{(3)(\text{aq})}$ species. *Radiochim Acta* 89, 511-518.

630 Bethke, C., 2008. *Geochemical and Biogeochemical Reaction Modeling*. Cambridge University Press,
631 pp. 565.

632 Black, L., Breen, C., Yarwood, J., Garbev, K., Stemmermann, P., Gasharova, B., 2007. Structural
633 features of C-S-H(I) and its carbonation in air - A Raman spectroscopic study. Part II: Carbonated
634 phases. *J Am Ceram Soc* 90, 908-917.

635 Bube, C., Metz, V., Schild, D., Rothe, J., Dardenne, K., Lagos, M., Plaschke, M., Kienzler, B., 2014.
636 Combining thermodynamic simulations, element and surface analytics to study U(VI) retention in
637 corroded cement monoliths upon > 20 years of leaching. *Phys Chem Earth* 70-71, 53-59.

638 Chang, J., Fang, Y., 2015. Quantitative analysis of accelerated carbonation products of the synthetic
639 calcium silicate hydrate(C-S-H) by QXRD and TG/MS. *J. Therm. Anal. Calorim.* 119, 57-62.

640 Chisholm-Brause, C.J., Berg, J.M., Little, K.M., Matzner, R.A., Morris, D.E., 2004. Uranyl sorption
641 by smectites: spectroscopic assessment of thermodynamic modeling. *J Colloid Interf Sci* 277, 366-
642 382.

643 Chisholm-Brause, C.J., Berg, J.M., Matzner, R.A., Morris, D.E., 2001. Uranium(VI) sorption
644 complexes on montmorillonite as a function of solution chemistry. *J Colloid Interf Sci* 233, 38-49.

645 Divet, L., Randriambololona, R., 1998. Delayed ettringite formation: The effect of temperature and
646 basicity on the interaction of sulphate and C-S-H phase. *Cement Concrete Res* 28, 357-363.

647 Elzinga, E.J., Tait, C.D., Reeder, R.J., Rector, K.D., Donohoe, R.J., Morris, D.E., 2004. Spectroscopic
648 investigation of U(VI) sorption at the calcite-water interface. *Geochim Cosmochim Acta* 68, 2437-
649 2448.

650 Fox, P.M., Davis, J.A., Zachara, J.M., 2006. The effect of calcium on aqueous uranium(VI) speciation
651 and adsorption to ferrihydrite and quartz. *Geochim Cosmochim Acta* 70, 1379-1387.

652 Gaona, X., Kulik, D.A., Macé, N., Wieland, E., 2012. Aqueous-solid solution thermodynamic model
653 of U(VI) uptake in C-S-H phases. *Appl Geochem* 27, 81-95.

654 Geipel, G., Reich, T., Brendler, V., Bernhard, G., Nitsche, H., 1997. Laser and X-ray spectroscopic
655 studies of uranium-calcite interface phenomena. *J Nucl Mater* 248, 408-411.

656 Grangeon, S., Claret, F., Linard, Y., Chiaberge, C., 2013. X-ray diffraction: a powerful tool to probe
657 and understand the structure of nanocrystalline calcium silicate hydrates. *Acta Crystallogr B* 69, 465-
658 473.

659 Hama, K., Kunimaru, T., Metcalfe, R., Martin, A.J., 2007. The hydrogeochemistry of argillaceous
660 rock formations at the Horonobe URL site, Japan. *Phys Chem Earth* 32, 170-180.

661 Harfouche, M., Wieland, E., Dähn, R., Fujita, T., Tits, J., Kunz, D., Tsukamoto, M., 2006. EXAFS
662 study of U(VI) uptake by calcium silicate hydrates. *J Colloid Interf Sci* 303, 195-204.

663 Hill, J., Harris, A.W., Manning, M., Chambers, A., Swanton, S.W., 2006. The effect of sodium
664 chloride on the dissolution of calcium silicate hydrate gels. *Waste Manage* 26, 758-768.

665 Ibanez, J., Artus, L., Cusco, R., Lopez, A., Menendez, E., 2007. Hydration and carbonation of
666 monoclinic C₂S and C₃S studied by Raman spectroscopy. *J. Raman Spectrosc.* 38, 61-67.

667 Joseph, C., Stockmann, M., Schmeide, K., Sachs, S., Brendler, V., Bernhard, G., 2013. Sorption of
668 U(VI) onto Opalinus Clay: Effects of pH and humic acid. *Appl Geochem* 36, 104-117.

669 Kienzler, B., Borkel, C., Metz, V., Schlieker, M., 2016. Long-Term Interactions of Full-Scale
670 Cemented Waste Simulates with Salt Brines. Karlsruhe Institut für Technologie (KIT) Report Nr.
671 7721.

672 Kienzler, B., Metz, V., Brendebach, B., Finck, N., Plaschke, M., Rabung, T., Rothe, J., Schild, D.,
673 2010. Chemical status of U(VI) in cemented waste forms under saline conditions. *Radiochim Acta* 98,
674 675-684.

675 Liu, G.K., Beitz, J.V., 2006. The Chemistry of the Actinide and Transactinide Elements. in: Morss,
676 L.R., Edelstein, N.M., Fuger, J. (Eds.). Springer, Dordrecht, The Netherlands.

677 Macé, N., Wieland, E., Dähn, R., Tits, J., Scheinost, A.C., 2013. EXAFS investigation on U(VI)
678 immobilization in hardened cement paste: influence of experimental conditions on speciation.
679 *Radiochim Acta* 101, 379-389.

680 Mazurek, M., 2004. Long-term Used Nuclear Fuel Waste Management – Geoscientific Review of the
681 Sedimentary Sequence in Southern Ontario. Institute of Geological Sciences, University of Bern,
682 Bern, Switzerland.

683 Noubactep, C., Sonnefeld, J., Merten, D., Heinrichs, T., Sauter, M., 2006. Effects of the presence of
684 pyrite and carbonate minerals on the kinetics of the uranium release from a natural rock. *J Radioanal*
685 *Nucl Chem* 270, 325-333.

686 OECD, 2006. Physics and Safety of Transmutation Systems: A Status Report. OECD Papers, 2 rue
687 Andre Pascal, 75775 Paris Cedex 16.

688 Peryt, T.M., Geluk, M.C., Mathiesen, A., Paul, J., Smith, K., 2010. Zechstein. in: Doornenbal, J.C.,
689 Stevenson, A.G. (Eds.). *Petroleum Geological Atlas of the Southern Permian Basin Area*. EAGE
690 Publications b.v., Houten, the Netherlands.

691 Pointeau, I., Landesman, C., Giffaut, E., Reiller, P., 2004. Reproducibility of the uptake of U(VI) onto
692 degraded cement pastes and calcium silicate hydrate phases. *Radiochim Acta* 92, 645-650.

693 Richardson, I.G., 2008. The calcium silicate hydrates. *Cement Concrete Res* 38, 137-158.

- 694 Rosenheim, A., Daehr, H., 1929. Uranium tetroxide-hydrate. *Z Anorg Allg Chem* 181, 177-182.
- 695 Sato, M., Matsuda, S., 1969. Structure of vaterite and infrared spectra. *Z Kristallogr Krist* 129, 405-
696 410.
- 697 Schmeide, K., Gürtler, S., Müller, K., Steudtner, R., Joseph, C., Bok, F., Brendler, V., 2014.
698 Interaction of U(VI) with Äspö diorite: A batch and in situ ATR FT-IR sorption study. *Appl Geochem*
699 49, 116-125.
- 700 Smith, K.F., Bryan, N.D., Swinburne, A.N., Bots, P., Shaw, S., Natrajan, L.S., Mosselmans, J.F.W.,
701 Livens, F.R., Morris, K., 2015. U(VI) behaviour in hyperalkaline calcite systems. *Geochim*
702 *Cosmochim Acta* 148, 343-359.
- 703 Stumpf, T., Tits, J., Walther, C., Wieland, E., Fanghänel, T., 2004. Uptake of trivalent actinides
704 (curium(III)) by hardened cement paste: a time-resolved laser fluorescence spectroscopy study. *J*
705 *Colloid Interf Sci* 276, 118-124.
- 706 Su, J., Wang, Z.M., Pan, D.Q., Li, J., 2014. Excited states and luminescent properties of UO_2F_2 and its
707 solvated complexes in aqueous solution. *Inorg Chem* 53, 7340-7350.
- 708 Takita, Y., Eto, M., Sugihara, H., Nagaoka, K., 2007. Promotion mechanism of co-existing NaCl in
709 the synthesis of CaCO_3 . *Mater Lett* 61, 3083-3085.
- 710 Taylor, H.F.W., 1993. Nanostructure of C-S-H - Current Status. *Adv Cem Based Mater* 1, 38-46.
- 711 Thiery, M., Dangla, P., Belin, P., Habert, G., Roussel, N., 2013. Carbonation kinetics of a bed of
712 recycled concrete aggregates: A laboratory study on model materials. *Cement Concrete Res* 46, 50-65.
- 713 Thoenen, T., Hummel, W., Berner, U., Curti, E., 2014. The PSI/Nagra Chemical Thermodynamic
714 Database 12/07.
- 715 Tits, J., Geipel, G., Macé, N., Eilzer, M., Wieland, E., 2011. Determination of uranium(VI) sorbed
716 species in calcium silicate hydrate phases: A laser-induced luminescence spectroscopy and batch
717 sorption study. *J Colloid Interf Sci* 359, 248-256.
- 718 Tits, J., Walther, C., Stumpf, T., Macé, N., Wieland, E., 2015. A luminescence line-narrowing
719 spectroscopic study of the uranium(VI) interaction with cementitious materials and titanium dioxide.
720 *Dalton Trans* 44, 966-976.
- 721 Torrents, A.M., 2014. Doctoral Thesis: Effect of Alkaline Conditions on Near-field Processes of a
722 Spent Nuclear Fuel Geological Repository. Department of Chemical Engineering. Universitat
723 Politècnica de Catalunya-Barcelona Tech, Catalunya.
- 724 Wang, Z.M., Zachara, J.M., Gassman, P.L., Liu, C.X., Qafoku, O., Yantasee, W., Catalano, J.G.,
725 2005a. Fluorescence spectroscopy of U(VI)-silicates and U(VI)-contaminated Hanford sediment.
726 *Geochim Cosmochim Acta* 69, 1391-1403.
- 727 Wang, Z.M., Zachara, J.M., McKinley, J.P., Smith, S.C., 2005b. Cryogenic laser induced U(VI)
728 fluorescence studies of a U(VI) substituted natural calcite: Implications to U(VI) speciation in
729 contaminated Hanford sediments. *Environ Sci Technol* 39, 2651-2659.
- 730 Wang, Z.M., Zachara, J.M., Yantasee, W., Gassman, P.L., Liu, C.X., Joly, A.G., 2004. Cryogenic
731 laser induced fluorescence characterization of U(VI) in Hanford vadose zone pore waters. *Environ Sci*
732 *Technol* 38, 5591-5597.
- 733 Wersin, P., Johnson, L.H., Schwyn, B., Berner, U., Curti, E., 2003. Redox Conditions in the Near-field
734 of a Repository for SF/HLW and ILW in Opalinus Clay. Nagra Technical Report NTB 02-13. Nagra,
735 Wettingen, Switzerland.
- 736 Wieland, E., Macé, N., Dähn, R., Kunz, D., Tits, J., 2010. Macro- and micro-scale studies on U(VI)
737 immobilization in hardened cement paste. *J Radioanal Nucl Chem* 286, 793-800.
- 738 Wieland, E., Tits, J., Spieler, P., Dobler, J.P., 1997. Interaction of Eu(III) and Th(IV) with sulphate-
739 resisting Portland cement. *Materials Research Society Symposium Proceedings*, pp. [d]573-578.

- 740 Wolfgramm, M., Thorwart, K., Rauppach, K., Brandes, J., 2011. Zusammensetzung, Herkunft und
741 Genese geothermaler Tiefengrundwässer im Norddeutschen Becken (NDB) und deren Relevanz für
742 die geothermische Nutzung. *Z. geol. Wiss.* 39, 173-193.
- 743 Yu, P., Kirkpatrick, R.J., Poe, B., McMillan, P.F., Cong, X.D., 1999. Structure of calcium silicate
744 hydrate (C-S-H): Near-, mid-, and far-infrared spectroscopy. *J Am Ceram Soc* 82, 742-748.
- 745 Zhao, P., Allen, P.G., Sylwester, E.R., Viani, B.E., 2000. The partitioning of uranium and neptunium
746 onto hydrothermally altered concrete. *Radiochim Acta* 88, 729-736.
- 747